A Hydrologic and Geochemical Investigation of Basin Infiltrated River Water in Alluvial, Basalt, and Combination Alluvial/Basalt Aquifers

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Abstract

Groundwater overdraft is a persistent problem in the western US including in the East Snake River Plain (ESRP) Aquifer (ESRPA), the largest in Idaho. Managed aquifer recharge (MAR) is a tool increasingly utilized for stabilizing aquifer storage volumes and is employed across the ESRP to infiltrate untreated river water through constructed recharge basins. Hydrologic and water quality data from monitoring well networks in the vicinity of three ESRP MAR basins – one each in an alluvial, basalt, and combination alluvial/basalt setting, together representative of the ESRPA – were examined. Processes investigated were hydrologic mixing between ambient groundwater and infiltrated water and whether cation exchange and calcite precipitation/dissolution reactions impacted MAR water. In addition, groundwater level increases and decreases in response to infiltration events were investigated as to whether they indicated the arrival and departure of MAR water at a pumping location. The results also provide suggestions regarding the location of hydrologic and chemical monitor well placement as MAR programs continue to evaluate new sites. A sample analysis provided results consistent with the stratification of infiltrated and ambient groundwater at all three sites. In the alluvial and basalt settings some sampling locations were entirely infiltrated water while at the combination alluvium/basalt setting vertical stratification in the aquifer was inferred. Hydrologic responses to basin inflow events in wells chemically unimpacted by MAR water indicated that water table responses did not necessarily evidence the arrival or departure of infiltrated water but could be the result of a "mound" of infiltrated water growing under the basins that displaced ambient groundwater away from it.

At the alluvial site an analysis of samples indicated that reverse cation exchange impacted infiltrated water, but calcite precipitation (if it occurred) did not impact infiltrated water chemistry. Geochemical modeling predicted that the reverse cation exchange processes would continue with increased alluvium interaction, but that calcite precipitation would have a nondetectable impact on the chemistry of infiltrated water and in any case would not be discernable from an analysis of the samples because much of the calcium precipitated as calcite would have been derived from reverse cation exchange. All wells, located between 620 ft (190 m) downgradient and 100 ft (30 m) cross gradient, were chemically impacted by infiltrated water (primary infiltration event average of 24.8 AF [30,590 m³] per day for 54 days) suggesting that these are appropriate distances for chemical monitor wells. However, local heterogeneities in the flow field played a larger role in the movement of infiltrated water than simple distance from the basin.

At the basalt site, fracture flow governed hydrologic and chemical impacts. Two distant (greater than 5,280 ft [1,609 m]) locations showed no chemical interactions between the aquifer matrix and infiltrated water. At a third location cation exchange processes may have impacted the samples. The three wells chemically impacted by infiltrated water were within 600 ft (183 m) of the basin while the other locations, 5,280 ft (1,609 m) or greater, were hydrologically impacted but showed no evidence of infiltrated water possibly because preferential flow paths routed the high volumes (primary infiltration event average of 730 AF [900,440 m³] per day for 133 days) away from the wells. Together these results suggest a range in which chemical monitor wells might be located but also demonstrate the significance of flow field heterogeneity.

At the combination alluvium/basalt site, wells were located 1,250 ft (380 m) cross gradient and 6,750 ft (2,060 m) downgradient and were not chemically impacted by the basin (infiltration event average of 88.3 AF [180,900 m³] per day for 136 days) suggesting that these locations are too distant to be appropriate for chemical monitoring. Hydrologic responses to infiltration events were observed only in the closer location.

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Disclaimer

The opinions and recommendations expressed in this work are those of the author and do not reflect those of the Idaho Water Resource Board, the Idaho Department of Water Resources, or the State of Idaho.

Chapter 1: Introduction

Groundwater overdraft is a persistent problem in the United States (US) with approximately 8.11 * 10⁸ acre-feet (AF) (1,000 km³) depleted between 1900 and 2008 [1] and is especially acute in the arid western portion of the country [2, 3, 4]. The East Snake River Plain (ESRP) Aquifer (ESRPA) is one of the aquifers experiencing overdraft [5, 6]. The ESRPA is in southern Idaho (Figure 1) and is one of the most productive aquifers in the US [7] as well as a sole source aquifer for more than 200,000 people [8]. The aquifer also supports the irrigation of approximately 1,237,000 acres (500,600 hectares) [6] and the largest commercial trout farming industry in the US relies on its discharges to the Snake River [9].



Figure 1 -- Location of the East Snake River Plain (ESRP) and its underlying aquifer (ESRPA) [10] in the State of Idaho as well as the Snake River, the approximate location of the Idaho National Laboratory (INL), and the approximate location of major springs in the Thousand Springs Complex discharging into the channel of the Snake River.

The ESRPA saw an increase in storage of approximately 17.5 million AF (21.6 km³) from 1912 to

1976, however, the aquifer has subsequently been depleted at an average rate of 0.32 million

AF/year (0.40 km³/year) (Figure 2). This depletion has been accompanied by substantial decreases in spring discharge to the Snake River resulting in, among other deleterious issues, potential widespread pumping curtailment across the ESRP as senior surface water right holders assert that junior groundwater users extract their water [13].



Figure 2 [11] – Increases and decreases of storage in the ESRPA and associated spring discharge rates at Thousand Springs Complex (which account for approximately 70% of the spring discharge from the ESRPA [calculated from 12]), 1912 – 2015.

Managed aquifer recharge (MAR) is a method of adding volume to the aquifer beyond that provided by seepage losses from rivers and incidental recharge from irrigation operations, thereby stabilizing and potentially increasing storage [14, 15, 16, 17] in arid watersheds. Globally, 1,200 MAR programs were in existence in 2015 [16] and increase in number by approximately five percent per year [17]. MAR strategies include the use of direct injection and recovery from the same or nearby wells (aquifer storage and recovery) of treated wastewater [15, 16] and untreated urban runoff [18, 19, 20], as well as the use of infiltration basins [15, 21, 22] utilizing water from a variety of sources (e.g., untreated river water).

Specific to the ESRPA, the State of Idaho through the Idaho Water Resource Board (IWRB) administers a MAR program (program) with a goal of recharging an average of 0.25 million acre feet (AF) (0.31 km³) annually [23, 24] across the ESRP. The program primarily utilizes canal seepage losses during the non-irrigation season as well as infiltration basins [22]. These infiltration basins source untreated water from the Snake River and provide high volumes (e.g., up to 1,200 AF [1.5 million m³] per day) of recharge in concentrated locations. This source of recharge is different from the distributed nature of recharge that provides most of the volume to the ESRPA and may alter the geochemical characteristics of the locally recharged groundwater because of concentrated infiltration and associated water-rock interactions. However, only a limited number of studies have addressed the hydrologic and geochemical evolution of MAR water in the vicinity of MAR infrastructure with the majority focusing on the geochemical evolution of water from various sources (e.g., pretreated wastewater [18, 25, 26] and urban stormwater runoff [19, 20, 21]) recharged via the aquifer storage and recovery method. No known studies examining hydrologic or geochemical processes resulting from water-rock interactions of untreated surface water in the vicinity of dedicated MAR infiltration basins were identified.

Here, to address the limited previous work, are the results of an investigation of hydrologic mixing, local groundwater level responses, and water-rock interactions influencing geochemical changes in groundwater chemistry during MAR operations in infiltration basins across the ESRP. Specifically, groundwater levels as well as infiltrating surface water and groundwater compositions in local monitor well networks were collected during recharge events at three MAR sites with differing geologic settings – alluvium, basalt, and combination alluvium/basalt, together characteristic of the ESRPA -- and used to infer process and their extent in influencing water composition. From this analysis, IWRB program staff and the public will be able to better interpret groundwater level data, including whether increases and decreases in groundwater levels indicate the arrival and departure of infiltrated water, and understand how MAR operations hydrologically and geochemically interact with the receiving substrate. The analysis will also be useful to the IWRB program as it continues to develop new infrastructure and associated monitor well networks in differing geologic settings across the ESRP.

East Snake River Plain Aquifer Hydrogeology

The ESRP is approximately 60 mile (100 km)-wide crescent-shaped plain which resulted from the North American continent passing west over a mantle hotspot now located under Yellowstone National Park [27] and the accompanying deposition of Pleistocene age (less than 2.6 Ma) basalt flows [28]. It is underlain by the ESRPA. Basin-and-range structures border the ESRP on its north, east, and south, the Island Park caldera to its northeast, and the Western Snake River Plain to the west [29]. More than 95% of the ESRPA host rock is basalt, with individual flows ranging in thickness from between 16.5 and 82 ft (5 and 25 m) and extending up to 30 miles (48 km) laterally [30]. Eruptions occurred intermittently, allowing for the intercalation of sedimentary layers between basalt flows [12]. The abundance of alluvial deposits increases at the outer boundaries of the aquifer [28] due to fluvial discharge from the surrounding mountains and accompanying sediment transport. Water movement through the aquifer is dominated by fracture flow within the basalt [31] and substantial (e.g., 5.8 million AF [7.15 km³] in 1939, 1965, 1976, 1984, and 1987 [Figure 2]) but declining quantities of water from the ESRPA discharge into the Snake River through a series of large springs [12, 32] including the Thousand Spring Complex shown in Figure 1.

Lower elevations of the ESRP receive approximately 10 inches (25 cm) of precipitation annually, while some headwater mountain catchments receive over 60 inches (150 cm) [32]. Numerous rivers, including the Snake and Big Lost Rivers, discharge into or flow across the ESRP and provide some recharge to the aquifer.

Large scale irrigation projects began on the northern and eastern portions of the ESRP in the 1880s and 1890s [33], and in the southern portion of the plain after the construction of the Milner Dam in 1902 [34]. Incidental losses from these irrigation projects, including canal seepage losses and the over-application of irrigation water, increased recharge to the aquifer by approximately 70% when compared to seepage from rivers [35], as shown in Figure 2. Starting in the 1950s, changes in irrigation practices, increased groundwater withdrawals, and persistent drought combined to decrease volumetric storage in the aquifer [32, 36, 37]. Specifically, irrigators began large-scale conversion to more efficient sprinkler systems while the use of turbine water pumps increased extractions as the technology was widely adopted in the post-WWII era. Concurrently, annual precipitation began declining in headwater catchments, leading to decreased river flows and associated recharge.

East Snake River Plain Aquifer Geochemistry

Investigations of water-rock interactions influencing ESRPA groundwater chemistry have largely focused on the data rich Idaho National Laboratory (INL) site (Figure 1). The field parameters and geochemical composition of the average (as calculated from [38]) non-contaminated groundwater sample from the upper 250 ft (76.2 m) of the ESRPA in the southeast portion of the INL (closest to the study areas) and surface waters sampled from tributary basins which supply incidental recharge to the aquifer at the INL site are provided in Table 1. The groundwaters are of a calcium-bicarbonate type, typically oversaturated with respect to carbonate minerals (aragonite, calcite, and dolomite)

and undersaturated with respect to gypsum. The typical sample collection date of the surface water in Table 1 was July 2, potentially impacting the reported concentrations and parameters (e.g., temperature).

Previous investigations of water-rock interactions within the ESRPA basalt minerals have been inferred the dissolution of olivine [40, 41, 42], pyroxene [40, 41], plagioclase [40, 42, 43, 44], pyrite [40, 42], potassium feldspar [41, 42, 43], silica [43], kaolinite [43], halite [42, 43], volcanic glass [44], and the congruent dissolution of olivine, diopside, amorphous silica, and anhydrite [45]. Precipitation of secondary calcite [40, 41, 42, 43, 44, 45], silica [40], montmorillonite [43], Caand/or Na-smectite [42], gibbsite [42], and ferric hydroxide [42, 45] has been observed or inferred. Other processes of potential relevance include cation exchange reactions [41, 44, 46] and the mixing of waters from differing recharge sources [38, 41, 42, 43]. Cation exchange capacities of ESRPA materials have variously been reported [46, 47, 48, 49, 50, 51, 52, 53]. These previous aquifer-scale water-rock interaction investigations provide results that are relevant at scales of tens to hundreds of miles (km) distance and years of time, much different than the distance and time scales characteristic of MAR operations which are localized to the vicinity of a recharge basin and occur on time scales of week to months.

In addition to aquifer-scale investigations, investigations of localized discharges and leaks of contaminated water associated with INL facilities and infrastructure at scales similar to MAR operations found that cation exchange processes, which occurred predominantly within sedimentary interbeds, influenced the migration of cations in the vadose zone and ESRPA [54, 55, 56, 57, 58].

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	Temperature (°C)	рН	Specific Conductance (µS/cm at 25 °C)	Alkalinity (mg/l as CaCO3)	Dissolved Oxygen (mg/l)	Ca ²⁺ (mg/l)	Mg ²⁺ (mg/l)	Na⁺ (mg/l)	K⁺ (mg/l)	HCO₃⁻ (mg/l)	
Average Tributary Surface Water	13.05	8.22	296.91	116.83	8.64	30.83	9.28	4.78	1.37	142.17	
Average Groundwater	14.29	8.12	365.4	137.5	7.7	35.89	13.73	15.11	3.26	167.6	
	CO₃²- (mg/l)	Cl ⁻ (mg/ l)	SO4 ²⁻ (mg/l)	F ⁻ (mg/l)	NO₃⁻ (mg/I)	SiO2 (mg/l)	Aragonite (log Q/K)	Calcite (log Q/K)	Dolomite (log Q/K)	Gypsum (log Q/K)	CO₂ (log fugacity)
Average Tributary Surface Water	1.16	3.28	14.74	0.18	0.09	10.70	0.25	0.42	1.38	-2.77	-3.14
Average Groundwater	0.964	17.6 8	17.63	0.53	1.13	30.23	0.28	0.45	1.55	-2.67	-2.97

Table 1 -- The geochemistry of average groundwater samples from the southeast portion of the INL and surface water tributary to the ERPA at the INL as calculated from [38]. Log Q/K and fCO₂ calculated via [39].

Chapter 2: Materials and Methods

Study Site Descriptions

Alluvial Site – The Jones Pit

Constructed in 2018, the four-acre (1.62 hectare) basin known as the Jones Pit is in the eastern portion of the ESRPA (Figure 3). Based on five recharge events between 2018 and 2021, the Jones Pit is capable of infiltrating approximately 30 AF (37,000 m³) per day [22]. The site is in Pleistocene era glacial outwash alluvium of the Snake River [33]. The alluvium is approximately 200 ft (61 m) thick and underlain by basalt **[59]**. Depth to water varies seasonally from a high of 50 ft (15 m) below ground surface (bgs) in October to a low of 110 ft (33.5 m) bgs in April [22]. Groundwater flow directions determined from the local potentiometric surface is from east to west (Figure 3) and is consistent with the regional (ESRPA) gradient provided by East Snake River Plain Aquifer Model (ESPAM) [10].



Figure 3 – The alluvial infiltration basin, known as the Jones Pit, and its monitor well network, as well as its location within ESRP. Callouts provide the sampling well names and approximate horizontal distances to the basin and direction of groundwater flow was provided by [10].

The monitoring network (Figure 3) consists of four wells located adjacent to the infiltration basin at distances of between 55 ft (17 m) and 620 ft (190 m). The IDWRMW was the only well with a pressure transducer for continuous monitoring of water levels and is located immediately

downgradient of the basin. The other monitoring wells, C. Arave, J. Stucki, and D. Stucki, are located increasingly farther away from the basin, orthogonal to the regional gradient. Wells logs are provided in Appendix A.

Water diverted from the South Fork of the Snake River below the Heise gauge (USGS monitoring location #13037500, two miles [3.2 km] to the east) is delivered to the basin by the Farmers Friend Irrigation Company (Farmers Friend). The Farmers Friend system is approximately 21.9 miles (35.2 km) from diversion to terminus and infiltrated approximately 140 AF (172,000 m³) per day during a 2018 infiltration event [22]. Assuming the canal seeps a uniform volume of water and that approximately 0.4 miles (0.64 km) of canal borders the vicinity of the site (Figure 3), 2.56 AF (3,160 m³) per day of incidental recharge probably occurs immediately adjacent to the basin. Assuming half flows towards the monitor well network, incidental recharge supplies approximately four percent of the volume when the basin is infiltrating its maximum capacity.

Basalt Site – MP29/31

Two subbasins in the western portion of the ESRPA, Milepost 29 and Milepost 31, connect on the surface during infiltration operations and can be considered as a single basin (MP29/31). Milepost 31 (MP31), an approximately 271-acre (100 hectare) basin, was constructed in 2014 and had subsequently been utilized annually during the non-irrigation season (i.e., September through June). Milepost 29 (MP29), an approximately 335-acre (135.6 hectare) basin, was constructed in 2020, and has been subsequently utilized in tandem with MP31. The basin has an operational capacity of approximately 1,200 AF (1.4 million m³) per day based on eight infiltration events [22]. The site is in an olivine tholeiite basalt section deposited approximately 2.4 Ma but includes some flows as old as 3 Ma [28] and extends beyond a depth of 363 ft (110.6 m) (Appendix B). A review of water table depths from 10/8/2019 through 10/8/2021 suggest that groundwater levels are controlled by events in the basin, rising during the non-irrigation season when infiltration operations occur and declining after their conclusion. Depth to water during that time ranged between 238.12 ft (72.58 m) bgs on 4/11/20 and 269.23 ft (82.06 m) bgs on 9/21/21. A 2016 fluorescein dye tracer test [60] (tracer test) found local east to west groundwater flow (Figure 4) consistent with the regional gradient provided by ESPAM, although no regionally-upgradient (upgradient) wells were sampled.



Figure 4 – The basalt infiltration basin(s), known as MP29/31, and its monitor well network, as well as its location within the ESRP. Callouts provide the sampling well names and approximate horizontal distances to the closest basin. Direction of groundwater flow provided by [10].

The monitoring network consists of five wells (Figure 4, well logs in Appendix B) each equipped with a pressure transducer allowing for continuous water level measurements. There are three close wells (East, Central, and MW1) located within 600 ft (183 m) of the basin and two outer wells (West and MW2) each located greater than a mile (1.6 km) from the basin. All wells are downgradient except for MW1, which is regionally cross-gradient (cross gradient) from both basins and 5,280 ft (1,609 m) from MP31.

Water diverted from the Snake River at Milner Dam is delivered by the American Falls Reservoir District Number 2 Irrigation District via the Milner-Gooding Canal to MP29/31. Headgates into the basin are located 29 and 31 miles (46.7 and 50 km) from the river. Between the river and the basin, the canal seeps approximately 59.5 AF (73,400 m³) per day [22]. Assuming uniform conveyance losses and 4.5 miles (7.2 km) of canal in the vicinity of the site (Figure 4), approximately 8.9 AF (1,100 m³) per day of canal seepage is estimated to impact the monitor network, or less than one percent of the inflow volume of the infiltration basin at maximum capacity.

Combination Alluvial/Basalt Site – West Egin Lake

The 300-acre (121 hectare) basin known as West Egin Lake is in the northern portion of the ESRPA (Figure 5). It is the largest basin in a privately operated recharge complex that is dispersed over an area of approximately 2.4 miles² (6.2 km²). Groundwater recharge in the complex began in 1972, after which the basin had been used consistently for MAR. West Egin Lake has a maximum measured infiltration capacity of approximately 220 AF (271,350 m³) per day [22]. The site is in an interlayered section of basalt and fine-grained sediments (well 07N 39E 07 BDA1 located on a knoll inside of the basin, Appendix C). The olivine tholeiite basalt consists of primarily of flow less than 2.6 Ma but includes some flows as old as 3 Ma [28] with a clay interbed identified at approximately 183 – 188 ft (55.8 - 57.3 m) bgs with basalt continuing beneath the interbed to a depth of at least 320 ft bgs (Appendix C). Depth to water ranged between 82.85 ft (25.25 m) and 84.90 ft (25.88 m) during the analyzed infiltration event.



Figure 5 – The combination alluvial/basalt infiltration basin (seen north of road and full of water), known as West Egin Lake, and its monitor well network, as well as its location within the ESRP. Callouts provide the sampling well names and approximate horizontal distances to the basin. A deep and a shallow well are drilled at each called out location. Direction of groundwater flow provided by [10].

Although the ESPAM predicted regional ground water flow from east to west (Figure 5) an observed correlation of increased streamflow in the Henry's Fork of the Snake River (located to the east and southeast of West Egin Lake) with MAR activities [61], suggested that the clay layer may serve as an aquitard, creating a local shallow aquifer that routed water back to the river rather than recharging the ESRPA thus minimizing the effectiveness of the MAR activities. In response, the Idaho Department of Water Resources (IDWR) drilled four monitor wells in 2020, one shallow and one deep in each of two locations (Figure 5, Appendix D includes the two deep well logs). The purpose of these wells were to determine the extent of the clay layer, whether the basin infiltrated into a local shallow aquifer that was distinct from the ESRPA, and to serve as dedicated water quality monitor wells for the site. A clay layer was observed at the West location, but not at the Easement location (Appendix D), suggesting that the ESRPA rather than a shallow local aquifer receives the infiltrated

water. The West Deep and Shallow wells produce water from below and above the clay layer, respectively. The West wells and the Easement Well Deep host pressure transducers. The Easement wells are slightly cross gradient from the infiltration basin and the West Wells are over a mile (1.6 km) down gradient from the basin.

Water diverted from the Henry's Fork of the Snake River near the municipality of St. Anthony is delivered to West Egin Lake via the St. Anthony Union Canal operated by Egin Bench Canal Company, Inc. The headgate into the basin is located approximately 14 miles (22.5 km) from the river. The St. Anthony Union Canal is approximately 20.5 miles (33 km) from diversion to terminus and infiltrated approximately 162.7 AF (200,700 m³) per day during a 2019 infiltration event [22]. A smaller basin, capable of infiltrating approximately 138.8 AF (171,200 m³) per day [22], is approximately two-miles (3.2 km) cross- and upgradient from West Egin Lake. Within that same two-mile (3.2 km) radius, approximately one mile (1.6 km) of canal is upgradient. Assuming uniform canal losses, half of which flow towards the basin and its monitor network, and that half of the volume infiltrated in the smaller basin flows towards West Egin Lake, approximately 73.5 AF (91,900 m³) per day of water infiltrated upgradient could hypothetically reach the monitor well network, or approximately one-third of the volume infiltrated through West Egin Lake at its maximum capacity.

Water Quality Data

Water samples and field parameters (pH, dissolved oxygen [DO], specific conductivity [SC], temperature) were collected from the infiltration basins and monitoring wells by IWRB program contractors. At the alluvium and basalt sites the samples were collected in compliance with sitespecific water quality monitoring regimes required by the Idaho Department of Environmental Quality (DEQ). At the combination alluvium and basalt site, which is not subject to DEQ monitoring, the samples in Table 7 were the only ones to have been collected from the monitor wells. Samples were collected in accordance with Standard Operating Procedures for the Statewide Groundwater Quality Monitoring Program [22]. Specifically, surface water grab samples were collected from just below the headgate or near it in the canal. For groundwater samples, wells were pumped until field parameters stabilized over three successive readings, each taken five minutes apart, or until three casing volumes of water had been removed. Field parameters were measured with probes calibrated daily. Water samples for major ions chemistry were collected, unfiltered, in 250 mL high density polyethylene bottles (triple rinsed with native water), leaving a small head space in the bottle for mixing (e.g., shaking) in the laboratory. Samples for cation and anion analysis were collected together and were not acid preserved. Samples were stored in an ice-filled cooler immediately after collection and shipped overnight for analysis to an Environmental Protection Agency certified laboratory in Boise, operated by the Idaho Bureau of Laboratories. Specific conductivity was measured using probes that were calibrated via a one-point method using a 500 μ S/cm solution. pH was measured with a glass electrode calibrated with buffer solutions with pH of 4.0, 7.0, and 10.0. DO meters were calibrated with the barometric pressure reported at the closest National Oceanic and Atmospheric Association reporting point (the Idaho Falls Regional Airport). Calcium, sodium, potassium, and magnesium concentrations were determined via inductively coupled plasma-atomic emission spectrometry [62]. Chloride and sulfate concentrations were determined via ion chromatography [63]. Alkalinity was determined via titration to an electrometrically determined end point of pH 4.5 [64]. Although measured as calcium carbonite alkalinity sampling results were returned as bicarbonate alkalinity, likely resulting from the assumption of neutral pH and dominance of bicarbonate, requiring that the results in the samples be multiplied by 1.219 (CaCO₃ mg/l to HCO₃⁻ mg/l) for reporting and analysis. Nitrate plus nitrite reported as N was determined via automated colorimetry [65].

Hydrologic Data

Inflows into the sites were measured in different ways. At the Jones Pit, a meter on an inlet pipe measured the pressure of water in the column, converted it to a rate in miner's inches averaged over 15-minute periods which allowed for volumes to be totalized into acre feet per day. The inflow rate was calibrated monthly via manual measurements of the canal flow rate above and below the inlet, performed by IWRB program staff using a SonTek Flowtracker 2 handheld Acoustic Doppler Velocimeter. Differences in reported inflow rates were interpolated between manual measurements. At MP29/31, inflows were measured using a weir and calibrated once by IWRB program staff during the analyzed recharge event with an Acoustic Doppler Current Profiler meter positioned on a moving boat. Measurements were within 1% of each other. At West Egin Lake, inflows were based on stage heights measured using a pressure transducer and a rating curve. Groundwater levels in monitor well networks were measured hourly using pressure transducers which were calibrated semiannually with a handheld electrical measurement tape. All transducers were gauged (vented to the surface) rendering barometric pressure correction unnecessary. Depth to water measurements were converted to water table elevation above the North American Vertical Datum of 1988 (NAVD 88). NAVD88 was calculated using either high accuracy GPS data collection or an automated process that digitally overlayed the latitude and longitude of the MP on a rectified topographic map and had an accuracy of \pm 10 ft (3.05 m).

Geochemical Modeling

Geochemical modeling was conducted using The Geochemist's Workbench® Release 16 software package (GWB) [39] and the associated *thermo* thermodynamic data base (based on the Lawrence Livermore National Laboratory (LLNL) dataset, EQ3/6 data file *data0.3245r46*) and ESRPA specific cation exchange selectivity coefficients [46].

Charge balance and saturation state of surface and groundwater samples with respect to aquifer minerals were calculated using the GWB SpecE8 module. Saturation indices (SI) are defined as:

$$SI = \log \frac{Q_{i,T}}{K_{i,T}} \tag{1}$$

where $Q_{i,T}$ is the ion activity product (or quotient) and $K_{i,T}$ is the equilibrium constant for *ith* mineral dissolution reaction at temperature *T*. The quotient is the numerical product of activities of the individual species appearing in the *ith* reaction raised to the power (positive for product and negative for reactants) of the stochiometric coefficient of that species in the reaction. The sign of *SI* indicates the saturation state of the water with respect to the mineral; supersaturated (positive), equilibrium (zero), or undersaturated (negative).

Mass transfer calculations [66] that involve geochemical reactions that transfer ions between aqueous and solid phases (e.g., cation exchange, precipitation/dissolution, etc.) were carried out using the React module of GWB. The modeling domain was configured as a continuously stirred tank reactor (CSTR) in which water-rock interactions are simulated in a system where infiltrated water mixes with and ultimately displaces groundwater. The water:rock ratio is calculated using a fully saturated infiltration-site specific porosity [67]. ESRPA relevant cation ion exchange capacities [46] were used to define rock reactive properties.

Chapter 3: Results and Discussion

Study-Wide Water Composition

Water samples were analyzed for approximately 120 analytes including field parameters, nutrients (nitrate plus nitrite as N and total phosphorus), bacteria (total coliform, E. coli), major ions (alkalinity reported as bicarbonate, calcium, chloride, magnesium, potassium, sodium, sulfate), and anthropogenic contaminants such as hydrocarbons (e.g., gasoline), pesticides (e.g., atrazine), and herbicides (e.g., chlorophenoxy acids). The results for the full suite of analytes are available [24]. No anthropogenic contaminant concentrations of regulatory concern were observed. The fifty-eight groundwater and nine surface water samples were analyzed in this investigation and their field parameters and major ion concentrations are reported in Tables 2, 5, 7. Charge imbalance errors ranged from +4.9% to -3.5% and were both positive and negative for all sites with no correlations noted. All waters were calcium-bicarbonate type except for the surface water at the combination alluvial/basalt site which was a sodium-bicarbonate type (Figure 6). The chemistry of water from the three sites were distinct from each other, reflecting the spatial variability of the ESRPA. The chemical variation between samples from individual sites was dominated by variations in anion concentrations (chiefly chloride and sulfate). All samples were oversaturated with respect to calcite, except one groundwater sample collected from the alluvial site (which had an anomalously low pH), and one groundwater sample and one surface water sample at the combination alluvial/basalt site.



Figure 6 -- Piper diagrams of all samples analyzed. 6a) Includes all groundwater samples. 6b) Includes all surface water samples.

Hydrochemical Conceptual Model of Basin Infiltrated Water

At each site, surface water is delivered to the infiltration basin from a canal through a headgate. Surface water in the basin vertically infiltrates through the vadose zone to the saturated zone where it mounds under the basin on top of ambient groundwater, as illustrated in Figure 7. Because rates [68] of evaporation are small compared to infiltrations rate, the effects of evaporation were ignored. Ambient groundwater is defined as aquifer water that is was not chemically impacted by the infiltration event. Mixing between ambient groundwater and infiltrated surface water, if it occurs, is at the infiltrated mound-ambient groundwater interface and is referred to as aquifer mixing. In addition, the pumping of hydrologically and impacted wells in which infiltrated water and groundwater are stratified during sampling will result in the mixing of the of the two waters. This type of mixing is referred to in-well mixing. The location of the mound remains centered under the basin, but because of its presence water infiltrated later in time can displace earlier infiltrated water in all directions away from the mound at a local scale (i.e, the scale of monitor well networks), including against the regional (i.e., ESRPA) gradient. The mounding in the absence of significant mixing results in a rise in the water table with water levels increasing in surrounding wells indicating the arrival of infiltrated water. The height of the infiltration mound under the basin increases and decreases as a function of infiltration rate.



Figure 7 -- Hypothesized hydrochemical conceptual model of basin-infiltrated water and its interaction with the receiving rock matrix.

As infiltrated water moves through the aquifer matrix it may be subject to water-rock interactions including cation exchange and dissolution/precipitate reactions. Because silicate mineral reaction rates in the ESRPA are slow [45] and the reaction rates for carbonate minerals are up to 1,000 times faster [69] only carbonate mineral (i.e., calcite) dissolution/precipitation reactions are considered important in the timeframe and scale of MAR operations.

Alluvial Site – The Jones Pit

Water Quality Sample Data

Water quality results for the 18 samples collected during the fall 2019 Jones Pit MAR operation are shown in Table 2. This operation was chosen for analysis because it represents the largest infiltration event. In addition, the water quality results (Figure 6) are similar to the four other MAR operations at the site. The fall 2019 infiltration event began on 8/23/2019. Groundwater samples collected before this date (8/21/2019) represent ambient conditions. Samples from each well and from the Headgate were collected on 9/27/2019 and 10/31/2019. A fourth sample was collected from all wells on 11/7/2019, seven days after the infiltration event ended.

Sample Location	Date	Electrical conductivity	рН	Temperature	O ₂ (aq)	Ca ²⁺	Na⁺	Mg ²⁺	K+	Cl⁻	SO4 ²⁻	HCO₃ ⁻	NO3 ⁻ + NO2 ⁻ as N	Charge imbalance error	Sodium Adsorption Ratio	Calcite
		uS/cm		С	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l			SI
C. Arave	8/21/2019	358	7.72	15.7	8.00	60	9.1	14.0	2.4	9.07	45.2	210.9	1.100	0.77%	0.27	0.35
C. Arave	9/27/2019	349	7.78	11.7	7.85	62	9.2	14.0	2.4	7.87	40.2	230.4	1.700	-0.23%	0.27	0.41
C. Arave	10/31/2019	325	7.37	8.4	9.16	52	8.8	12.0	2.2	10.20	40.7	181.6	0.410	2.81%	0.29	-0.24
C. Arave	11/7/2019	323	7.76	11.1	8.77	52	8.5	12.0	2.2	10.20	40.3	181.6	0.360	0.20%	0.28	0.21
J. Stucki	8/21/2019	378	7.62	15.9		66	9.4	16.0	2.6	8.20	40.8	245.0	2.000	1.67%	0.27	0.35
J. Stucki	9/27/2019	403	7.74	13.9	7.44	69	9.5	15.0	2.5	7.33	37.6	252.3	2.600	1.48%	0.27	0.47
J. Stucki	10/31/2019	327	7.84	7.7	9.10	53	8.6	12.0	2.2	10.50	40.6	179.2	0.320	1.00%	0.28	0.25
J. Stucki	11/7/2019	325	7.73	11.2	12.55	52	8.3	11.0	2.1	11.40	41.8	175.5	0.190	-0.47%	0.27	0.17
D Stucki	8/21/2019	360	7.70	18.2		60	9.6	13.0	2.0	7.88	39.1	213.3	0.450	1.54%	0.29	0.37
D Stucki	9/27/2019	367	7.82	12.5	5.31	64	9.7	13.0	1.9	6.91	35.0	242.6	0.480	-0.56%	0.29	0.49
D Stucki	10/31/2019	318	7.84	7.1	7.05	53	8.9	11.0	1.8	11.10	41.2	175.5	0.084	0.43%	0.29	0.23
D Stucki	11/7/2019	325	7.86	10.6	10.79	52	8.8	11.0	1.8	11.70	42.0	175.5	0.076	-0.91%	0.29	0.29
IDWRMW	8/21/2019	362	7.80	18.0		68	10.0	13.0	1.8	7.91	40.1	229.2	2.100	2.22%	0.29	0.55
IDWRMW	9/27/2019	289	8.00	11.5	6.16	50	9.0	9.3	1.6	9.08	38.9	156.0	0.120	1.62%	0.31	0.39
IDWRMW	10/31/2019	334	7.85	10.6	8.87	55	11.0	10.0	1.8	15.50	49.5	163.3	0.010	0.37%	0.36	0.27
IDWRMW	11/7/2019	334	7.93	11.9	9.05	54	11.0	10.0	1.7	15.40	49.2	164.6	0.010	-0.74%	0.36	0.37
Headgate	9/27/2019	299	8.50	12.0	6.67	45	10.0	11.0	1.7	10.70	40.5	156.0	0.009	-1.67%	0.35	0.84
Headgate	10/31/2019	367	8.57	4.0	9.90	55	14.0	14.0	2.2	17.10	53.7	182.8	0.009	-0.98%	0.44	0.93

Table 2 – Water quality results from the alluvial site – The Jones Pit, calcite SI calculated by [39].

Hydrogeologic Investigation

Figure 8 shows the aquifer water table elevation measured in the IDWRMW, inflow volumes of infiltrating water into the basin, and an overview of the Anderson Canal operations. The MAR operation occurred as two events, the first and smaller (approximately 10 AF [12,335 m³] per day) from 8/24/2019 to 8/27/2019 and the second larger (approximately 25 AF (38,000 m³] per day) from 9/8/2019 to 10/31/2019 which together infiltrated 1,380 AF (1.7 million m³) of surface water. The Anderson Canal is located approximately 1,100 ft (335 m) south of the Jones Pit and diverted for irrigation operations 500 AF (616,740 m³) per day or more from the Snake River during the month of August [70] (High Canal Flows). On 9/1/2019 the canal began a near linear reduction in its diversion rate that lasted until 9/24/2019 (Declining Canal Flows), when it stabilized at approximately 100 AF (123,350 m³) per day (Low Canal Flows) until concluding irrigation operations on 10/9/2019 (Figure 8).



Figure 8 -- Inflows into the alluvial site -- Jones Pit, water levels at the IDWRMW, and a general description of nearby canal operations vs time.

Groundwater levels (Figure 8) reflect the combined effect of MAR operations and variable incidental recharge from the Anderson Canal. The rise in water table elevation prior to the larger MAR event corresponds to high canal flows. The primary infiltration operation began on 9/8/2019 near the start of declining canal flows and resulted in a six day rise in groundwater elevation that began on the same day as the larger infiltration operation and therefore was probably not the arrival of infiltrated

water at the location. Following this rise, the groundwater level remained relatively constant until the flow in the Anderson Canal was stopped, at which time the groundwater level declined even though the infiltration rate remained constant. When MAR operations were completed on 10/31/2019 the groundwater levels rapidly decreased.

Sample Source Determination and Hydrologic Mixing Analysis

The chloride, sulfate, and nitrate plus nitrite concentrations for groundwater samples collected on 8/21/2019 were distinct from the headgate samples collected on 9/27/2019 and 10/31/2019 (Table 2) suggesting that these anions could be used as tracers to assess mixing of infiltrated surface water(s) and ambient groundwater. Chloride and sulfate are naturally present in water due to the dissolution of ocean evaporates at the start of the hydrologic cycle (e.g., NaCl [halite] and CaSO₄ [anhydrite]) and can also be from anthropogenic sources (i.e., KCl and CaSO₄·2H₂O [agricultural fertilizers]) [47]. Because of its low affinity for anion exchange [71] chloride is commonly used as a conservative (i.e., non-reactive) aqueous tracer in groundwater studies [72, 73 74] on the temporal scale of weeks relevant to MAR events. In addition, under oxidizing conditions and low concentrations characteristic of the ESRPA, sulfate can also serve as a conservative aqueous tracer as demonstrated by its strong correlation (not shown) with chloride concentration. Because chloride and sulfate are correlated, the sulfate-to-chloride ratio (ratio) was utilized to estimate the extent of mixing between ambient groundwater and infiltrated surface water. Nitrate across the ESRP is primarily sourced from inorganic fertilizer applied to agricultural fields [47] and is also conservative on the timescales of MAR events.

Examining the sulfate-to-chloride ratios and nitrate concentrations of all groundwater and headgate water samples (Figure 9) shows that the groundwater samples fall into two groups, those interpreted as ambient groundwater (all samples from 8/21/2019 and samples from the C. Arave, D. Stucki, and D. Stucki wells [more distant wells] from 9/27/2019) with relatively high ratios and nitrate concentrations, and those interpreted as a mixture of infiltrated water and ambient groundwater (IDWRMW [closest well] from 9/27 and all samples from 10/31 and 11/7/2019) with ratios and nitrate concentrations intermediate between ambient groundwater and headgate water. The groundwater level was declining when chemically impacted samples were collected from the IDWRMW (Figure 8) indicating that water table decreases do not necessarily demonstrate the departure of infiltrated water.



Figure 9 -- Sulfate-to-chloride ratios vs nitrate concentrations in all samples at the alluvial site -- Jones Pit.

Although there are two sources of water, the composition of the headgate varied as function of time (Figure 9). In addition to the 9/27/2019 and 10/31/2019 headgate water quality data, weekly SC measurements between 9/12/2019 and 11/7/2019 (Figure 10, Appendix E) suggest that the surface water composition varied systematically and continuously. This trend may reflect a decreasing fraction of water released from Palisades Reservoir during the MAR event and an increasing fraction of relatively stable base flows resulting in total decreased flows at the Heise gauge [70] (Figure 10). Regardless of the reason for the trend, surface water composition at time between sampling event can be treated as a mixture of the 9/27/2019 and 10/31/2019 headgate water.



Figure 10 -- Specific conductance at the Headgate and flow of the Snake River at the Heise gauge vs date at the alluvial site – Jones Pit.

To estimate the fractions of ambient groundwater and infiltrated headgate water in the mixed groundwater sample, a three-end member mixing model (the ambient groundwater composition of each well [outer well samples from 9/27/2019 and IDWRMW sample from 8/21/2019], and the 9/27/2019 and 10/31/2019 headgate compositions) was used. The three-endmember mixing equation is

$$[X_i]_{mixed} = f_1[X_i]_1 + f_2[X_i]_2 + f_3[X_i]_3$$
(2)
Where

$$f_3 = 1 - f_1 - f_2 \tag{3}$$

and $[X_i]_{mixed}$ is the concentration of the *ith* conserved ion in the mixed groundwater, f_1 , f_2 , and f_3 are the mixing fraction of the indexed end members, and $[X_i]_1$, $[X_i]_2$, and $[X_i]_3$ are the concentrations of *ith* conserved ion in the indexed end member resulting in three mixing equations, one each for sulfate, chloride, and nitrate. The Solver function of Microsoft Excel [75] was used to estimate f_1 and f_2 by minimizing the sum of squares of the relative difference between the measure and calculated sample concentrations.

$$\sum \left(\frac{[X_i]_{measured} - [X_i]_{calculated}}{[X_i]_{measured}}\right)^2 \tag{4}$$

The estimated mixing fractions are given in Table 3. As expected, a decrease in the fraction of ambient groundwater with increased duration of the infiltration event was observed for all wells. Complete displacement of ambient groundwater by infiltrated water is inferred in the 55 ft (17 m) regionally downgradient (downgradient) IDWRMW on 10/31/2019 and 11/7/2019 and indicates that the increased water table elevation beginning on the day of the infiltration operation might evidence the growth of a mound of infiltrated water under the basin which transmitted ambient groundwater away from it. The limited mixing of infiltrated water with ambient groundwater observed in the location on 9/27/2019 is consistent with a stratified aquifer with infiltrated water "sitting" atop the ambient groundwater and mixing of the two waters occurring in the well during pumping.

Table 3 – The estimated percentage of water from ambient groundwater (sample collected from each location on 9/27/2019 [8/21/2019 IDWRMW]), water entering the basin on 9/27/2019 (H 9/27), and water entering the basin on 10/31/2019 (H 10/31) for each groundwater sample collected on 9/27/2019, 10/31/2019, and 11/7/2019 at the alluvial site – Jones Pit.

		IDWRMW	/	D. Stucki				
	H 9/27 H 10/31 Ambie			H 9/27	7 H 10/31 Ambier			
9/27/2019	94.7%	0.0%	5.3%	0.0%	0.0%	100.0%		
10/31/2019	24.9%	75.0%	0.0%	68.4%	15.7%	15.9%		
11/7/2019	26.5%	73.5%	0.0%	61.7%	24.1%	14.2%		
		J. Stucki		C. Arave				
	H 9/27	H 10/31	Ambient	H 9/27	H 10/31	Ambient		
9/27/2019	0.0%	0.0%	100.0%	0.0%	0.0%	100.0%		
10/31/2019	84.8%	3.2%	12.0%	73.6%	2.7%	23.7%		
11/7/2019	78.3%	14.7%	7.0%	77.9%	1.4%	20.8%		

No evidence of infiltrated water was observed in the more distant wells on 9/27/2019 (Table 3). On 10/31/2019, the D. Stucki well, which is downgradient and farthest (620 ft [190m]) from the basin, was 16% ambient groundwater compared to 12% for the closer (200 ft [61 m]) and the slightly orthogonal J. Stucki well. The cross-gradient C. Arave well is located 100 ft (30 m) from the basin and had the largest percentage of ambient water remaining in its sample on 10/31/2019. The chemical presence of infiltrated water at cross-gradient wells further suggests the presence of a mound under the basin that allows infiltrated water to move across the regional gradient. The results also indicate that local heterogeneities in the flow field play a larger role in the movement of water during the infiltration event than simple distance from the basin. By 11/7/2019, the percentage of ambient groundwater in all of the more distant wells had decreased compared to 10/31/2019, consistent with the infiltrated water moving through the monitoring wells between sampling events and seven

days after the infiltration event concluded. In addition, and as expected, the relative abundance of 10/31/2019 compared to 9/27/2019 headgate water in the D. Stucki and J. Stucki wells increased as newer headgate water displaced older headgate water. At the C. Arave the opposite was observed, although for both the 10/31/2019 and 11/7/2019 samples the fraction of 10/31/2019 headgate water is estimated to be less than 3%. These results from wells that are cased to approximately the same depth near the bottom of the alluvium (Appendix A) indicate that the infiltrated water packet may have been elongated and moved faster in the direction of the regional gradient than orthogonal to it, and that ambient water is displaced from the top-down.

Travel times to the monitoring wells can be bounded by the mixing fractions estimates. The larger infiltration event began on 9/8/2019 and by 9/27/2019 95% of the water at IDWRMW was estimated to be headgate water suggesting a travel time of significantly less than 19 days. No infiltrated water was observed in the more distant wells on 9/27/2019 but was on 10/31/2019 suggesting travel times of between 20 and 53 days, with relative travel times being J Stucki $\leq D$ Stucki < C Arave.

Geochemical Investigation

In addition to mixing, infiltrated waters may undergo water-rock interactions with the aquifer matrix through cation exchange reactions (Equation 5) and/or dissolution-precipitation of carbonate minerals (i.e., calcite, Equation 6).

$$>X:Na^{+} + 0.5Ca^{2+} \leftrightarrow Na^{+} + >X:Ca^{2+}_{0.5}$$
(5)

 $CaCO_3 + CO_2 + H_2O \leftrightarrow Ca^{2+} + 2HCO_3^{--}$

These potential reactions were assessed by using the mixing fractions from Table 3 and the end member water compositions from Table 2 to calculate the devised compositions of mixed waters in the absence of any other processes. The devised compositions were compared to the measured compositions of the corresponding groundwater sample. The difference between the devised and measured compositions are referred to as [Species]_{excess} and are shown in Table 4. Positive and negative values of [Species]_{excess} indicate the addition or removal of a species from the groundwater by water-rock interactions, respectively. The estimated uncertainties (standard deviation) in these calculations are 0.09, 0.09, 0.04, 0.01, and 0.13 meq/L for calcium, magnesium, sodium, potassium, and bicarbonate, respectively, as described in Appendix F.

(6)

			[Sp	ecies] _{Excess} lor	n Concentrations					
			IDWRMW		C Arave					
		9/27/2019	10/31/2019	11/7/2019	9/27/2019	10/31/2019	11/7/2019			
Ca ²⁺	meq/L	0.19	0.12	0.08		0.14	0.17			
Mg ²⁺	meq/L	-0.15	-0.27	-0.26		0.02	0.03			
Na⁺	meq/L	-0.04	-0.09	-0.08		-0.05	-0.06			
K+	meq/L	0.00	-0.01	-0.01		0.01	0.01			
∑Cation _{excess}	meq/L	-0.01	-0.24	-0.28		0.11	0.14			
HCO3 [−]	meq/L	0.00	-0.09	-0.10		0.44	0.27			
			J Stucki		D Stucki					
Ca ²⁺	meq/L		0.24	0.19		0.17	0.09			
Mg ²⁺	meq/L		0.03	-0.06		-0.06	-0.08			
Na ⁺	meq/L		-0.06	-0.10		-0.07	-0.09			
K ⁺	meq/L		0.01	0.01		0.00	0.00			
∑Cation _{excess}	meq/L		0.22	0.04		0.03	-0.08			
HCO₃ ⁻	meq/L		0.27	0.28		0.12	0.10			

Table 4 – [Species]_{Excess} ion concentrations in chemically impacted samples at the alluvial site -- Jones Pit.

As may be seen in Table 4 many of the [Species]_{excess} values are of similar or smaller magnitude than their respective uncertainties (especially at the two-standard deviations level) that make it difficult to draw definitive conclusions for individual sampling locations and dates. This uncertainty is highlighted by the differences between $\Sigma Cation_{excess}$ and HCO_3^- [Species]_{Excess} values which because of charge balance requirements should be the same ([Species]_{Excess} values for Cl⁻ and SO₄²⁻ are essentially zero as they were used to calculate the mixing fractions). However, by averaging across all locations and dates, [Species]_{Excess} values of 0.15, -0.09, -0.07, 0.00 and 0.14 meg/L for calcium, magnesium, sodium, potassium, and bicarbonate, respectively are calculated and can be used to infer specific geochemical processes. The almost equal milliequivalent increase in calcium and decrease in combined magnesium and sodium is consistent with cation exchange reactions in which magnesium and sodium in the infiltrated water displace calcium from aquifer sediments. Alternatively, the equal milliequivalent increase in calcium and bicarbonate is consistent with the dissolution of calcite. However, because all groundwater samples are oversaturated with respect to calcite (Table 2, C Arave 10/31/2019 being an outlier) dissolution of calcite is highly unlikely. Cation exchange with exchange site preference of Na⁺ \approx Mg²⁺ > Ca²⁺ was the only geochemical process that can be confirmed during the MAR operation.

To further assess the geochemical interactions a mass transfer model using the GWB React module was developed that accounted for coupled cation exchange and calcite precipitation. Conceptually

ambient groundwater, represented by the average composition of all 8/24/2019 and 9/27/2019 (except IDWRMW) groundwater samples, in a fixed volume of pore space was displaced by four pore volumes of infiltrating 10/31/2019 Headgate water resulting in less than 1% of ambient water remaining at the end of the simulation. The volume (and mass) of aquifer sediments were estimated using a porosity of 0.413 (the average of ESRPA sedimentary interbeds [67]) and a sediment grain density of 2.65 g/cm³ resulting in a water:rock ratio of 1:3.8. A cation exchange capacity (CEC) of 0.21 meq/g_{rock} and Gapon exchange coefficients measured for ESRP sediments [46] were used to describe equilibrium cation exchange. Calcite precipitation was modeled using a mixed kineticequilibrium approach. On average ESRPA groundwaters are oversaturated with respect to calcite (Table 1, SI of approximately 0.4), however, widespread rapid calcite precipitation is not observed suggesting that at this level of oversaturation precipitation is very slow in comparison to the duration of MAR operations. To account for this in the model calcite was only allowed to precipitate if its SI value was greater than 0.4. Six different reaction scenarios were considered:

- 1. Mixing (no cation exchange or calcite precipitation),
- 2. Mixing and precipitation (no cation exchange),
- 3. Mixing and cation exchange (no calcite precipitation),
- 4. Mixing, cation exchange, and calcite precipitation,
- Mixing and cation exchange (no calcite precipitation) with a 10x increase in the CEC (equivalent to a volume of water seeing more aquifer sediments),
- 6. Mixing, cation exchange, and calcite precipitation with a 10x increase in the CEC.

Simulation results (Appendix G) are presented in Figure 11 in terms of the sodium adsorption ratio (SAR) [76] and the sulfate-to-chloride ratio. Also shown in Figure 11 are values for Jones Pit waters. SAR is a derived parameter to assess irrigation water quality and is defined as:

$$SAR = \frac{[Na^+]}{\sqrt{\frac{[Ca^{2+}] + [Mg^{2+}]}{2}}}$$
(7)

where concentrations are in meq/L. Because SAR includes contributions from the major cations, it is a simplified proxy for assessing water-rock interaction.


Figure 11 – A graphical depiction of modeling results for various scenarios with mixing functions (sulfate-to-chloride ratios) vs sodium adsorption ratio (SAR) functions at the alluvial site -- Jones Pit.

The sulfate-to-chloride ratios in Figure 11 demonstrate that mixing (Scenario 1) is predicted to play a primary role in the evolution of the infiltrated water along with cation exchange (Scenarios 3 and 5), which dominates the changes in SAR values. The precipitation of calcite (Scenarios 2, 4, and 6) is predicted to have minimal effect on SAR values. Two samples from the IDWRMW plot close to the Scenario 3 curve indicating that cation exchange might have impacted samples collected at this well and they also plot in the vicinity of the Scenario 4 curve suggesting that calcite precipitation might also have occurred but its impact on SAR was nondetectable. The SAR of samples from the outer wells plot near the Scenarios 5 and 6 curves with less scatter than ambient groundwater indicating that cation exchange reactions might have continued to impact the infiltrated water as it interacted with additional sediments in the longer flow path to the outer wells. Again, the precipitation of calcite is predicted to have minimal to no impact on the SAR values for these wells. Simulated changes in aquifer sediments for Scenario 6 (mixing, cation exchange, and calcite precipitation) are presented in Figure 12 and given in Appendix H. The figure shows that as infiltrated water displaces ambient groundwater it promotes cation exchange in which of sodium and magnesium in the mixed groundwater displace calcium from cation exchange sites in equal equivalent concentrations, a result that is consistent with the direct observations based on [Species]_{excess} calculations presented earlier. In addition, a majority (70 to 80%) of the desorbed

calcium is precipitated as calcite rather than remaining in solution. This explains why calcite precipitation, if it is occurring, cannot be concluded solely from groundwater measurements. Overall, the coupled cation exchange – calcite precipitation reaction is given by $Na^{+} + 0.5Mg^{2+} + 2>X:Ca^{2+}_{0.5} + 2HCO_{3}^{-} \rightarrow CaCO_{3} + H_{2}O + CO_{2} + >X:Na^{+} + >X:Mg^{+}_{0.5}$ and supports the exchange site preference of $Na^{+} \approx Mg^{2+} > Ca^{2+}$ inferred from Table 4.



Figure 12 – Modeling predictions for the change in cation concentrations on exchange sites and calcite precipitated vs infiltrated water in the system at the alluvial site -- Jones Pit.

ESRP sediments have cation selectivity of $Ca^{2+} > Mg^{2+} > K^+ > Na^+$ [44, 45] probably because bonds formed with divalent cations are stronger than those formed with monovalent cations. However, reverse cation exchange [77] occurs when monovalent sodium is removed from water by a stochiometric exchange of divalent calcium and/or magnesium from the aquifer matrix despite the generally higher affinities of exchange sites for the latter. Reverse cation change can occur because, in addition to exchange site affinity, cation exchange processes are influenced by the distribution of cations available for exchange and therefore are related to the activity of cations in solution. As the activity of sodium, potassium, and/or magnesium increases relative to calcium in water, the affinity of exchange sites for the monovalent cation or magnesium can exceed the affinity for calcium [44] promoting reverse cation exchange.

(8)

Synthesis and Conceptual Model

Water levels in the vicinity of the Jones Pit basin were controlled by the nearby Anderson Canal and the MAR event in the basin. Two of the three samples from the closest well were 100% infiltrated water and all hydrologic mixing observed was inferred to have occurred in-well. All wells, located between 100 ft (30 m) cross gradient and 620 ft (190 m) downgradient, were impacted by infiltrated water suggesting that these are appropriate distances for chemical monitor wells in alluvium. However, local heterogeneities in the flow field played a larger role in the movement of infiltrated water during the MAR event than simple distance from the basin. Water level rises did not necessarily indicate the arrival of infiltrated water but might have evidenced the growth of a mound of infiltrated water under the basin that transmitted ambient groundwater away from it. In addition, water table declines did not indicate the departure of infiltrated water. Reverse cation exchange was observed and was modeled to be the primary process impacting infiltrated water and predicted to continue with increased interactions with sediments. The precipitation of calcite was not found to be an important process influencing water composition. The findings are synthesized in a reinterpretation of Figure 7 and presented in Figure 13. Figure 13 is not to scale but is best suited for the 10/31/2019 sampling date when MAR was still occurring and a mound of infiltrated water was present under the basin.



Figure 13 – A conceptual model of the alluvial site -- Jones Pit that is not to scale but is most applicable to the 10/31/2019 sampling date.

Basalt Site – MP29/31

Water Quality Sample Data

Water quality results for 41 samples-collected at the MP29/31 site between 8/24/2020 and 4/8/2021 (sampling season) in association with the two events are shown in Table 6. Field parameters for the six surface water samples were not measured as part of sampling and instead temporally relevant measurements at USGS gauge number 13093383, Snake River at Pigeon Cove near Twin Falls, Idaho located 22.5 miles (36.25 km) southwest of the site are provided. The fall of 2020 through spring of 2021 operation was chosen for analysis because it was the most recent operation at the commencement of this investigation. In addition, water quality data from the seven other MP29/31 MAR operations were similar to those shown Figure 6. There were two primary recharge events, one at MP31 occurring from 9/8/2020 through 9/28/2020 and a second in both basins from 11/6/2020 through 3/18/2021.

Sample Location	Date	Electrical Conductivity	рН	Temper- ature	O₂(aq)	Ca ²⁺	Na⁺	Mg ²⁺	K+	Cl-	SO4 ²⁻	HCO3-	NO3 ⁻ + NO2 ⁻ as N	Charge Imbalance Error	Sodium Adsorption Ratio	Cal- cite
		uS/cm		С	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l			SI
East Well	8/24/2020	435.4	7.96	9.3	8.1	40	24	15	4.1	27.3	40.1	171.9	1.05	0.03%	0.82	0.25
East Well	9/17/2020	410.9	7.88	10.4	5.9	40	18	14	3.8	18.5	38.7	164.6	0.432	0.94%	0.62	0.17
East Well	10/7/2020	428.9	7.85	13	5.9	41	20	15	4.2	19.8	38.8	170.7	0.519	2.08%	0.68	0.20
East Well	1/6/2021	541	8.3	8.5	9.8	47	27	18	6.0	32.8	53.4	209.7	2.01	-3.52%	0.85	0.71
East Well	2/9/2021	540	8.24	5.6	9.1	47	29	19	6.6	33.9	52.8	208.4	1.56	-1.38%	0.90	0.61
East Well	3/11/2021	447.9	8.14	5.3	7.7	48	29	19	6.5	34.6	51.3	199.9	1.72	0.73%	0.90	0.50
East Well	4/8/2021	444.8	7.82	6.2	8.3	47	32	18		35.0	53.3	196.3	1.6	-0.04%	1.01	0.17
Central Well	8/24/2020					41	24	15	4.2	26.4	40.6	176.8	0.963	-0.83%	0.82	
Central Well	9/17/2020	437.2	7.81	9.7	7	43	20	17	4.2	20.1	40.3	175.5	0.895	3.90%	0.65	0.14
Central Well	10/7/2020	438	7.88	10.5	6.1	42	20	15	4.3	19.8	38.6	180.4	9.48	-0.94%	0.67	0.23
Central Well	1/6/2021	540	7.84	10.9	7.9	47	27	18	5.3	32.6	53.8	208.4	2.08	-2.46%	0.85	0.28
Central Well	2/9/2021	546	8.01	8.7	8.9	48	28	19	5.5	33.8	53.2	210.9	1.63	-1.60%	0.87	0.44
Central Well	3/11/2021	445.2	7.94	7.6	8.8	49	29	19	5.7	34.0	50.7	202.4	1.54	1.36%	0.89	0.34
Central Well	4/8/2021	433.6	7.86	7.9	6.2	46	29	18	5.4	33.9	51.7	188.9	1.55	1.22%	0.92	0.21
West Well	8/24/2020	435.6	7.96	14.5	7.7	40	21	17	4.2	22.9	40.7	185.3	0.883	-0.84%	0.70	0.35
West Well	9/17/2020	434.8	7.92	14.6	8.8	40	20	16	4.1	22.3	39.2	173.1	0.839	0.66%	0.68	0.29
West Well	10/7/2020	432.8	8	14.5	8.9	39	21	16	4.2	21.5	37.8	173.1	0.829	1.04%	0.72	0.36
West Well	1/6/2021	432.8	7.97	14.1	8.2	38	20	16	4.2	23.2	40.7	178	0.916	-2.31%	0.69	0.32
West Well	2/9/2021	434.4	7.99	14.2	8.4	37	21	16	4.3	23.0	39.6	178	0.87	-2.06%	0.73	0.33

Table 5 -- Water quality results from the basalt site – MP29/31, calcite SI calculated by [39].

Sample Location	Date	Electrical Conductivity	рН	Temperature	O₂(aq)	Ca ²⁺	Na⁺	Mg ²⁺	K⁺	Cl-	SO4 ²⁻	HCO₃ ⁻	NO3 ⁻ + NO2 ⁻ as N	Charge Imbalance Error	Sodium Adsorption Ratio	Calcite
		uS/cm		С	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l			SI
West Well	3/11/2021	362.8	7.89	14.1	7.9	39	21	17	4.3	22.4	38.8	178	0.839	0.84%	0.71	0.25
West Well	4/8/2021	358.6	7.85	14.4	8.7	39	21	16	4.2	22.9	39.8	174.3	0.836	0.23%	0.72	0.21
MW1	10/7/2020	432.2	8.03	14.3	8.6	41	22	14	3.6	23.4	37.6	169.4	0.826	0.65%	0.76	0.40
MW1	11/16/2020	470.7	8.3	13.7	8.9	45	27	17	4.0	34.3	52.2	185.3	1.43	-2.18%	0.87	0.72
MW1	12/8/2020	531	8.15	11.2	9	46	30	18	4.8	32.8	53.7	197.5	1.51	-0.79%	0.95	0.57
MW1	1/6/2021	555	7.99	7.3	7.8	47	31	19	5.4	35.9	56	209.7	1.93	-1.73%	0.96	0.38
MW1	2/9/2021	560	8	8	8.6	48	31	19	5.0	35.7	54.2	212.1	1.88	-1.34%	0.96	0.42
MW1	3/11/2021	446.8	8.08	7.4	9	49	31	18	5.2	36.6	51.7	198.7	1.58	0.60%	0.96	0.47
MW1	4/8/2021	431	7.91	7.8	7.4	46	30	17	5.0	34.7	51.5	191.4	1.5	-0.01%	0.96	0.27
MW2	10/7/2020	382.4	8.1	13.9	9.1	36	18	13	3.4	19.3	33.6	149.9	0.58	0.69%	0.65	0.37
MW2	11/16/2020	381.5	8.27	13.8	8.7	34	18	13	3.4	19.6	34	152.4	0.617	-1.90%	0.67	0.52
MW2	12/8/2020	377.8	8.08	13.7	9.1	35	18	14	3.4	20.0	34.9	154.8	0.644	-0.57%	0.65	0.34
MW2	1/6/2021	378.1	8.12	13.6	9.2	34	17	13	3.3	20.4	35.6	152.4	0.664	-3.00%	0.63	0.36
MW2	2/9/2021	378.8	8.13	13.7	8.9	34	18	13	3.4	20.4	34.5	154.8	0.642	-2.58%	0.67	0.38
MW2	3/11/2021	314.8	8.02	13.7	8.7	35	18	14	3.4	20.0	35	154.8	0.62	-0.45%	0.65	0.28
MW2	4/8/2021	315	8	13.8	9	36	18	14	3.4	20.0	34.8	153.6	0.61	0.61%	0.64	0.27
Headgate	9/17/2020	614	8.3	13.5	10.4	43	18	15	3.4	18.1	38	157.3	0.611	4.49%	0.60	0.65
Headgate	11/16/2020	656	8.3	10.1	11.5	42	26	18	6.3	27.8	49.1	178	0.96	1.35%	0.85	0.63
Headgate	12/8/2020	692	8.3	7	12.1	45	28	19	7.0	31.4	54	181.6	1.32	2.03%	0.88	0.62
Headgate	1/6/2021	673	8.5	8	11.8	48	27	19	6.6	33.0	53.8	197.5	1.78	-0.62%	0.83	0.88
Headgate	2/9/2021	670	8.6	8.1	12.4	46	29	19	7.8	33.8	52.1	192.6	1.48	0.31%	0.91	0.95
Headgate	3/11/2021	658	8.7	11.3	12.9	47	30	19	8.4	34.9	51.4	169.4	1.43	4.89%	0.93	1.05

Table 5 -- Water quality results from the basalt site – MP29/31 continued.

Sample Source Determination

The sulfate-to-chloride ratios for all samples over the duration of the sampling season (Figure 14) show that groundwater samples from the West Well and MW2 (outer wells) are similar, groundwater samples from the Central Well and East Wells (inner wells) are similar but different from the outer wells, and groundwater samples from cross gradient MW1 are different from both. These three groups will be discussed individually.





The sulfate-to-chloride ratios measured in surface water decreased almost linearly during MAR operations. After an initial increase from ambient pre-infiltration values, sulfate-to-chloride ratios for the inner wells were nearly identical to and show similar decreases with time as the infiltrating surface water. This observation is consistent with the rapid and complete displacement of ambient water with infiltrated surface water. In contrast sulfate-to-chloride ratios in the outer wells remained constant during the entire infiltration event suggesting that infiltrated water did not reach the locations. In addition, the ratios for outer wells differ from the pre-infiltration values for the inner wells showing the spatial heterogeneity in water composition.

A fluorescence dye tracer test conducted at MP31 in 2016 during an infiltration event found an average groundwater velocity of 445 ft (135.6 m) per day [60]. Based on this velocity infiltrated water would be expected to reach the inner wells (200 ft [61 m] or less down gradient) within a day

and the outer wells (5,930 ft [1,810 m] to 6,775 ft [2,065 m] down gradient) within two weeks. For the inner wells, this estimated travel time is consistent with the changes in sulfate-to-chloride ratios. However, for the outer wells the presence of infiltrated water was not chemically observed indicating that these wells are not within the flow field of the MAR operation. This discrepancy may be the result of preferential fracture flow characteristic of the ESRPA basalts. The sulfate-to-chloride ratio at MW1 did not respond to the September MP31 infiltration event as the 10/7/2020 value was like the pre-infiltration inner wells but distinctly lower than that in the

the 10/7/2020 value was like the pre-infiltration inner wells but distinctly lower than that in the surface water and the outer wells (Figure 14). Because MW1 is located 5,280 ft (1,609 m) and cross gradient from MP31, this is not unexpected. However, when infiltration began in the much closer MP29 (600 ft [183 m] cross gradient) the sulfate-to-chloride ratio for MW1 paralleled those of the surface water and the inner wells albeit at a lower value indicating chemical impact at the location from MP29 only.

Collectively these results show that the inner wells chemically respond to infiltration at MP31 and probably MP29, MW1 only chemically responds to infiltration at MP29, and the outer wells do not chemically respond to infiltration in either basin highlighting the flow heterogeneity of the site.

Hydrogeologic Investigation

The water levels in all wells and inflows into both MP29 and MP31 are shown in Figure 15. To facilitate comparison among wells groundwater levels are relative to the sampling season minimums at each location. The MAR event in September infiltrated approximately 6,634 AF (8.2 million m³) of surface water at MP31 and the larger event from November through March infiltrated approximately 64,775 AF (80 million m³) at MP29 and 32,750 AF (40 million m³) at MP31.



Figure 15 -- Inflows into the basalt site – MP29/31 and water levels in all wells in the monitor well network vs time.

Prior to the September infiltration event, groundwater levels had been declining (Figure 15) at the location since the completion of a 78,355 AF [96.7 million m³] infiltration event in the spring of 2020 [24]. With the initiation of infiltration at MP31 the water table at the inner wells show up to a four ft (1.2 m) rise as a result of the arrival of infiltrated surface water (Figures 14 and 15). In addition, the water table rises to lesser extents at MW1 and the outer wells, locations that had not seen chemical evidence of infiltrated surface water (Figure 14), indicating that water level increases and decreases do not necessarily correspond to the arrival and departure of infiltrated water at a location. Instead, these hydrologic responses, when coupled with the complete displacement of ambient groundwater inferred in the inner wells above, might evidence the growth of a mound of infiltrated water directly under the basin which transmitted ambient groundwater away from it.

Figure 15 shows that the water table elevation of MW1 is heavily influenced by events in MP29. The September infiltration event utilized only MP31, and the hydrologic response at MW1 was similar to the West Well, 5,930 ft (1,808 m) downgradient from MP31. By contrast, in response to the early November inflow peak into MP29 MW1 responded similarly to the East Well, located a similar

distance (535 ft [163 m] cross gradient) from MP29. This observation coupled with the parallel variations in the sulfate-to-chloride ratio of infiltrated and MW1 water (Figure 14) suggest that although not wholly displacing ambient groundwater (as was the case for the inner wells) some amount of infiltrated water is likely present at the MW1 location when MP29 is operating. Comparison of the water levels in MW1 with the East well during the late December basin switch demonstrates that the East Well is more hydrologically influenced by events in MP31 than the more distant MP29.

Geochemical Investigation

Table 6 provides the average and standard deviation of the concentration difference between the

groundwater and Headgate water samples collected on the same.

Table 6 – Average concentration difference between groundwater samples at two locations, the East and Central wells (inner wells) and MW1, and the Headgate water as well as the standard deviations among the samples at the basalt site – MP29/31.

	Inner	Wells	M١	W1
	Average	St. Dev	Average	St. Dev
	me	q/L	me	q/L
Ca ²⁺	0.01	0.09	0.07	0.08
Na⁺	-0.01	0.04	0.09	0.05
Mg ²⁺	-0.01	0.08	-0.05	0.04
K+	-0.03	0.03	-0.06	0.02
Cl	0.00	0.02	0.08	0.06
SO4 ²⁻	0.01	0.02	0.03	0.05

For the inner wells all ion concentrations differences are indistinct from zero (Table 6) and indicate that the water in these wells was surface water chemically unaltered by cation exchange or precipitation reactions despite being oversaturated with respect to calcite (Table 5). The apparent lack of water-rock reaction may be the result of low cation exchange capacity of basalt, the short residence time, or both.

Groundwater sampled at MW1 during the November to March infiltration event had chloride and sulfate concentrations higher than the infiltrating surface water and the pre-infiltration MW1 groundwater (Table 6) suggesting that there may be additional source of the ions such as road salt (NaCl) and soil amendments (e.g., CaSO₄·2H₂O). The observed (Table 6) positive equal milliequivalent sodium, chloride, calcium, and sulfate differences are consistent with these possible ion sources. Negative differences of potassium and magnesium suggest that cation exchange may be occurring.

Figure 16 shows sulfate-to-chloride and sodium adsorption ratios for MW1 groundwater and infiltrated surface water. Not shown on the figure are the ratios for East and Central wells which are similar to the surface water. The SAR value for MW1 shifts from its pre-MP29 infiltration values of 0.76 to 0.87 10 days after the start of infiltration to a near constant more sodium rich value of 0.96 one month after the start of infiltration. The shift in SAR during infiltration contrasts to the near constant values observed for the Jones Pit wells J. Stucki, D. Stucki, and C. Arave and indicate that unlike the alluvium at the Jones Pit the basalt does not buffer the cation composition of the water. However, the constant SAR values and decreasing sulfate-to-chloride ratios after 12/8/2020 are consistent with some amount of buffering by cation exchange (maybe due to interactions with some sediments) occurring only when MP29 is in use. The hydrological and geochemical contrast between MW1 and the inner wells highlight the importance of local heterogeneity in controlling the hydrochemistry of fractured basalt.





Because the outer wells had near constant chemical composition and were chemically uninfluenced by the infiltration event, they were used to assess the combined uncertainties associated with repeated sampling and chemically characterizing groundwater wells over time, the result of which are given in Appendix F.

Synthesis and Conceptual Model

Distance from the basins influenced the water level responses in wells while flow heterogeneities (i.e., preferential fracture flow) controlled mixing and chemical impacts. The inner wells respond hydrologically and chemically to relatively close MP31. The water table at the East well specifically responds more to infiltration at MP31 than to MP29 but the chemical impact of MP29 in isolation on the inner wells could not be determined. Samples from the inner wells were completely infiltrated water demonstrating that displacement occurs without observable mixing and no chemical reactions were inferred. By contrast the water levels at MW1 were heavily influenced by the closer MP29 and only chemically impacted by events in that basin. Samples collected from MW1 showed some evidence of cation exchange only when MP29 was operating, demonstrating the importance of heterogeneous fracture flow. The chemically impacted wells were all within 600 ft (183 m) of the basin(s) which contrasts to the outer wells and MW1 when only MP29 was in operation, which were all greater than 5,280 ft (1,609 m) from the basin(s) and demonstrated no chemical evidence of infiltrated water despite the large infiltration event. This was probably because flow heterogeneities routed the infiltrated water away from the locations. Nevertheless, relatively muted water table changes in response to infiltration events were observed in the farther away locations and demonstrate that water table rises and falls in response to infiltration events do not necessarily indicate the arrival and departure of infiltrated water. Because of the dominance of fracture flow Figure 7 was not appropriate and instead Figure 4 was reinterpreted as a conceptual model of the site when both MP29 and MP31 are being utilized and presented in Figure 17.



Figure 17 -- A conceptual model of MP29/31 when both basins are operating.

Combination Alluvial/Basalt Site -- West Egin Lake

Water Quality Sample Data

West Egin Lake water quality results for the seven samples collected from four dedicated monitoring wells and two samples of surface water associated with the 7/1 - 11/13/2021 MAR operations are given in Table 8 and shown in Figure 6. Samples were collected in the spring of 2021 (prior to infiltration) from both the deep and shallow West wells, the Easement Well Deep, and (because an infiltration operation was not occurring) from a canal sourced from the Henry's Fork of the Snake River approximately seven miles (11.3 km) away from the basin. Samples were collected again during the summer from all four wells and at the basin 34 days after the infiltration event began.

Sample ID	Date	рН	Temperature	O ₂ (aq)	Ca ²⁺	Na⁺	Mg ²⁺	K⁺	Cl⁻	SO4 ²⁻	HCO3-	NO3 ⁻ + NO2 ⁻ as N	Charge imbalance error	Sodium Adsorption Ratio	Calcite
			С	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l			SI
Surface Spring	3/2/21	8.61	3.2	31	12.94	16.69	3.785	2.402	6.032	4.5	84.96	0.25	-1.92%	0.55	0.10
West Deep Well Spring	3/2/21	7.94	10.8	16.3	28.02	11.26	7.465	2.468	5.163	7.227	131.7	1.44	1.63%	0.49	0.04
West Shallow Well Spring	4/13/21	7.76	11.8	14.32	29.99	11.38	7.974	2.493	5.196	7.603	139	1.56	2.71%	0.48	-0.09
Easement Well Deep Spring	3/2/21	8.07	9.8	15.56	31	11	8.3	2.4	5.47	9.63	140.2	2.20	1.13%	0.45	0.22
West Shallow Well Summer	8/4/21	8.06	17.5	14.86	29.43	11.52	7.954	2.526	5.203	7.948	141.4	1.62	0.09%	0.49	0.30
West Deep Well Summer	8/4/21	8.01	19.3	13.58	28.95	11.5	7.753	2.526	5.215	8.171	141.4	1.69	-0.74%	0.49	0.27
Easement Well Deep Summer	8/4/21	7.94	15.7	14.23	30.04	11.22	8.341	2.44	5.406	9.432	142.6	2.08	0.31%	0.47	0.16
Easement Well Shallow Summer	8/4/21	7.48	16	15.74	28.81	12.35	5.467	2.505	4.584	4.467	139	0.98	1.56%	0.55	-0.33
Surface Summer	8/4/21	8.18	23.9	17.71	12.07	15.01	3.748	2.422	5.516	3.828	82.89	0.17	-2.68%	0.51	-0.07

Table 7 – Water quality results from the combination alluvial/basalt site – West Egin Lake, calcite SI calculated by [39].

Sample Source Determination

Chloride and sulfate concentrations for West Egin Lake waters are shown in Figure 18. Samples collected before and during the infiltration from the deep and shallow West location have similar chloride and sulfate concentrations as does the deep well at the Easement location (the shallow well was not sampled prior to the start of the infiltration event). Collectively, these groundwater samples define a linear trend between chloride and sulfate suggesting similar processes influence water composition. These samples are considered ambient groundwater. However, the surface water samples are not part of the trend defined by the groundwater samples. Samples from the Deep and Shallow wells at the West location have similar sulfate and chloride concentration suggesting the presence of a sedimentary interbed at this location does not alter the groundwater chemistry. However, at the Easement location the sample from the Shallow and Deep wells represent the extremes of the groundwater trend shown in Figure 18 demonstrating that vertical stratification of groundwater composition can occur in the absence of a sedimentary interbed. Because the shallow well was only sampled during the infiltration event, whether the stratification pre-dated the start of infiltration is not known.



Figure 18 -- Sulfate vs. chloride concentrations in all samples at the combination alluvial/basalt site -- West Egin Lake.

Hydrogeologic Investigation

The water levels in all wells and inflows into the West Egin Lake infiltration basin are shown in Figure 19. To facilitate comparison among wells groundwater levels are relative to the sampling season minimums at each location. The infiltration event occurred from 7/1 through 11/13/2021 and recharged 12,012.5 AF (14.8 million m³).



Figure 19 -- Inflows into the combination alluvial/basalt site – West Egin Lake and water levels in the monitor well network vs time.

Prior to the infiltration event, groundwater levels had been declining (Figure 19) at the location since the completion of a 18,275 AF [22.54 million m³] infiltration event in the Fall of 2020 [22]. Because the volume of the examined infiltration event was small, and the West location is over 6,750 ft (2,060 m) from the basin it is likely that the major upward trend beginning in late July was an increase in the regional water table elevation. Because the hydrologic impact of the infiltration event was not clearly observed in the West location it was likely on the order of the difference between the West location and the Easement Well Deep which began diverging on 7/23/2021, with a maximum difference of 0.28 ft (0.09 m) on 8/12/2021 and then converged on 9/11/2021 until diverging again in response to the peak inflow event in October. A sample that was not a mixture of ambient groundwater and sampled surface water was collected from the Easement Well Deep on 8/3/2021 indicating that water table rises do not necessarily indicate the arrival of infiltrated water at a location.

Geochemical Investigation

Table 8 provides the change in ion concentrations for the same well between the spring and summer sampling events (Temporal Changes by Season) and the difference in concentrations between the deep and shallow wells at the same location from the same sampling event (Vertical Change by Location).

Table 8 -	– Temporal	and vertical	differences	among select i	on concentra	ations in all sa	amples at the	combination
alluvial/	basalt site -	- West Egin L	.ake.					

	Те	mporal Cha	anges by Season (m	eq/L)	Vertical	Change by Loc	ation (meq/L)
	V	Vest	Eacomont Doon	Surface	١	Vest	Easement
lon	Deep	Shallow	Easement Deep	Suitace	Spring	Summer	Summer
Ca ²⁺	-0.05	0.03	0.05	0.04	-0.10	-0.02	0.06
Na ⁺	-0.01	-0.01	-0.01	0.07	-0.01	0.00	-0.05
Mg ²⁺	-0.02	0.00	0.00	0.00	-0.04	-0.02	0.24
K⁺	0.00	0.00	0.00	0.00	0.00	0.00	0.00
Cl-	0.00	0.00	0.00	0.01	0.00	0.00	0.02
SO4 ²⁻	-0.02	-0.01	0.00	0.01	-0.01	0.00	0.10

Table 8 shows little temporal differences in ion concentrations for the three wells suggesting that groundwater chemistry was unchanged by MAR operations. Vertically, there is also little difference in the composition of shallow and deep groundwater for the West location indicating chemical homogeneity of the aquifer despite the presence of a local sedimentary interbed.

By contrast, at the Easement location groundwater was vertically stratified during the summer sampling event, with shallow water being depleted with respect to magnesium, calcium, sulfate, and chloride, and enriched in sodium compared to the deep well. To further explore the variations in water chemistry, a simple mixing model was constructed with the two most extreme water samples (Easement Well Deep – spring and Easement Well Shallow – summer). The mixing model results are given in Figure 20 and show that all the groundwater samples can be described as a mixture of the two endmembers. In addition, the summer surface water sample also falls on the mixing curve (the spring surface water sample location was seven miles [11.3 km] from the infiltration basin). The surface water sample falls between the summer Easement Well Shallow and any other groundwater compositions indicating that the shallow Easement groundwater cannot be a mixture of the infiltrated surface water and ambient groundwater. However, because the surface and groundwater fall along the same mixing line, they are likely both subject to the same, unidentified hydrochemical process(es).



Figure 20 – Sulfate-to-chloride ratio vs SAR in all samples at West Egin Lake -- combination alluvial/basalt site with a mixing line between the Easement Well Deep – Spring and Summer samples.

Synthesis

The region of the aquifer chemically impacted by the infiltration event did not include any of the wells and hydrologic responses were not evident in the West location 6,750 ft (2,060 m) downgradient probably because of the relatively low infiltration rate and distance to the monitoring wells. The West location has a sedimentary interbed between the deep and shallow wells while the Easement location does not, but groundwater stratification was only observed in the latter indicating that a cohesive aquifer underlies the basin in which vertical stratification occurs. A sample from the Easement location was taken when the water table was elevated in response to the infiltration event but was not a mixture of the surface water and the other groundwaters demonstrating that rising water tables in response to infiltration events do not necessarily indicate the arrival of infiltrated water.

Chapter 4: Conclusions

This study was conducted to investigate hydrologic mixing, local groundwater level responses, and water-rock interactions influencing geochemical changes in groundwater chemistry during MAR operations in infiltrations across the ESRP. In pursuit of these goals, groundwater level and water quality sample data from three MAR basins, one each in alluvium, basalt, and combination alluvium/basalt, were evaluated. Together these locations are representative of the geology of the ESRPA. The major findings of this study are:

Hydrologic mixing and groundwater level data:

- Complete displacement of ambient groundwater by infiltrated water was observed in some wells at the alluvial and basalt sites. In addition, water table increases and decreases did not necessarily indicate the arrival and departure of infiltrated water at a given location but rather provide evidence of the growth of a mound of infiltrated water under the basin which displaced ambient groundwater.
- At the combination alluvial/basalt site there was a discontinuous sedimentary interbed and the chemistries of water above and below it were the same, but chemical stratification occurred at a second location absent the interbed. Water table increases and decreases in that location did not evidence the arrival and departure of infiltrated water.

Geochemical processes impacting infiltrated water:

- An analysis of water quality samples indicated that a type of reverse cation exchange impacted the infiltrated water at the alluvial site in which a sediment preference of Na⁺ ≈ Mg²⁺ > Ca²⁺ was inferred. Geochemical modeling predicted that the reactions were possible and would increase with additional sediment interaction.
- The precipitation of calcite would have minimal impact on the chemical evolution of infiltrated water.
- At the basalt site, no observable reactions impacted infiltrated water at two relatively close downgradient wells, but cation exchange processes probably impacted infiltrated water at a cross gradient well only when the closest basin to it was operating, highlighting the importance of heterogeneous fracture flow.

Additional findings synthesized with general applications for monitor well network designs:

Hydrologic monitoring wells: All monitor wells at the basalt site, ranging between 100 ft (30 m) and 6,775 ft (2,065 m) from the basin, responded hydrologically to the substantial inflow events in the basin with locations closer to the basin showing larger water level increases. At the

combination alluvial/basalt site the location 1,250 ft (380 m) from the basin responded hydrologically to the relatively small inflow event while wells 6,750 ft (2,060 m) from the basin apparently did not. Unsurprisingly, these results indicate that the hydrologic monitor well network designs should consider both volumes to-be infiltrated and distance from the basin.

Chemical monitoring wells: At the alluvial site, an infiltrated water packet may have been elongated, moving faster in the direction of the regional gradient than orthogonal to it, and all wells, which were located between 100 ft (30 m) cross gradient and 620 ft (190 m) downgradient, were chemically impacted by the basin. At the basalt site, evidence of infiltrated water was observed at locations 600 ft (183 m) or less from the basin but not at locations 5,280 ft (1,609 m) or greater probably because preferential flow paths routed the large volume of infiltrated water away from the more distant locations. At the combination alluvial/basalt site, no chemical evidence of infiltrated water was observed in a well 1,250 ft (380 m) from the basin. Together these results offer some range on the distance of chemical impacts of the basins but more broadly demonstrate that local heterogeneities in the flow field should be considered when designing chemical monitor well networks.

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B. CAS	ING/LIN	ER:	1												t
Diameter (nominal)	From (ft)	To (ft)	Gauge/ Schedule	Mate	rial Casing	Liner Threaded	Welded	-							
6	+2	158	,25	steel			×							-	┝
															t
								-						-	
									-			000	FIVES	-	┝
Vas driv	ve shoe i	used? 🔀	Y DN	Shoe De	epth(s) 15	8						HEG	EIVED		F
. PERF	ORATI	ONS/SC	REENS:		r							MAR	2 8 2019		
Perforati	ions 🖪	Y DN	Method _	per	twater					-		Department of	Water Resource	6	⊢
Aanufac	tured sc	reen 📙	YDANT	ype _	-							Easta	m Region		t
weithica (or install	ation		Diameter	THE CT	1	_								
rom (II)	To (ft)	Slot size	Number/ft	(nominal)	Material	Gauge or Sc	hedule	Comple	ted Dept	h (Measur	able):	158	-		
22	140	18,2	12	6	Steel	125		Date St	arted	3-12	-18	Date Comp	leted: 3-13	-19	
		-						14. DR	ILLER'S	S CERTI	FICATIO	N:			
ength o	f Headpi	pe		Lenat	h of Tailpine			the time	e the rig	was remo	oved.	onstruction stand	ards were compa-	ad with a	at
acker [UY DE	N Type		a				Compa	ny Name	. Der	nins	Drilling	Co. No.	79	
0.FILT	ER PAC	K:						*Princin	al Drille	_	J.Y.	E S	Data TR	201	1
Filter	Material	From	1 (ft) To	(ft) Qua	antity (lbs or fl ³)	Placement m	bodie	e moq	An Drine		any .	C SCIMULA	Late PI	201	-
								*Driller	-	11	1	1	Date	1.7.1	~
								*Operal		the	Da	rgen	Date_3-	13-1	1
1. FLO	WING A	RTESIA	N:					Operato	or I	-		~	Date		_

Appendix A – Alluvial Site Well Logs

IDWRMW

Druze WELL DRILLER'S REPORT 2.5 Inspected by	-
WELL TAG NO. D OOZ 66666 Twp Rge_Sec_ DRILLING PERMIT NO.	
DRILLING PERMIT NO.	
Water Right or Injection Well No. IZ. WELL IESIS: Lat. Long: : Pump Bailer Air Flowing Artesian Vield gal/min. Drawdown Pumping Level Time tame_Delos Stuck? Time Time	_
2. OWNER: Name Delos Stuck? Time	
Vame Delos Stucki	
uddress 121 x1 N. 95E.	
Xity tislaho talls State ID Zip 83401	
B. LOCATION OF WELL by legal description: Water TempBottom hole temp.	
/ou must provide address or Lot, Blk, Sub. or Directions to well. Water Quality test or comments:	_
wp. 3 North 15 or South Depth first Water Encounter	
ige. 3.9 East K or West D On Cloud LOG: (Describe repars or abandonment)	Vate
And the second s	1
at: :: Long: :::	
Address of Well Site 12181 N. 95 E. 2' 13' Clay	
(She if ket and of out - Dance in Bad of Lecture) City Liberto Falls IS Growel & Clay	
1 Bik Sub. Name 100 / 031 Class	-
Kelling La Children C	
10 10 Gravely Dond, Gib ot clay A	+
to Ge.	+
Differential Injection Other	+
TYPE OF WORK check all that apply (Replacement etc.) Anew Well Modify Abandonment Other	+
Child Internation Cable Mud Rotary Other Seal Material From To Weight / Volume Seal Placement Method	
Vertante o' 20'300# Ourbore	+
Vas drive shoe used? SV N Shoe Denth(s)	-
As drive shoe seal tested? Y SLN How?	
	-
CADING/LINER: Second	+-
6" +2' 178', 25 steel & K	-
	÷
	+-
ength of Headpipe Length of Tailpipe	
acker UY UN type JUN 2.5 2003	
PERFORATIONS/SCREENS PACKER TYPE	
erforation Method Department of Water Resources	-
creen Type & Method of Installation	+
rium to star size Number Diameter Material Gasing Liner Completed Depth 178/ Measure	able
	2010
Date: started <u>61210</u> Completed <u>613</u>	1
0 FILTER PACK	ho
I/We centry that all minimum well construction standards were consided with at a	ie.
Filter Material From To Weight / Volume Placement Method time the rig was removed.	1.0
Filter Material From To Weight / Volume Placement Method Filter Material From To Weight / Volume Placement Method time the rig was removed. Company Name Company Name Company Name Company Name Company Name	19
Filter Material From To Weight / Volume Placement Method Filter Material From To Weight / Volume Placement Method 1. STATIC WATER LEVEL OR ARTESIAN PRESSURE: Principal Drillow John Jack Smn Jack Date 6 / 3 / 4	23
Filter Material From To Weight / Volume Placement Method Filter Material From To Weight / Volume Placement Method I. STATIC WATER LEVEL OR ARTESIAN PRESSURE: Company Name Company Name Filter Material I. STATIC WATER LEVEL OR ARTESIAN PRESSURE: Principal Driller Company Name Simular Company Material Artesian pressure Ib, Date 6/3/4	23
Filter Material From To Weight / Volume Placement Method I. STATIC WATER LEVEL OR ARTESIAN PRESSURE: Company Name Company Name Company Name 1. STATIC WATER LEVEL OR ARTESIAN PRESSURE: Principal Driller Company Name Company Name St. t. below ground Artesian pressure Ib, Principal Driller Company Name Sint Company Name Difference It. Describe access port or control devices: Difference Date 6/13/0	22
Filter Material From To Weight / Volume Placement Method I. STATIC WATER LEVEL OR ARTESIAN PRESSURE: Company Name Statistic material of the placement Method LS_ft. below ground Artesian pressureib, Ib, principal Driller or Operator II Statistic material Date Operator I	23
Filter Material From To Weight / Volume Placement Method Instantial From To Weight / Volume Placement Method Instantial From To Weight / Volume Placement Method Instantic Water Level OR ARTESIAN PRESSURE: Company Name Image: Company Name Image: Company Name St. ft. below ground Artesian pressure Ib. Bb. Principal Driller Compartor II Date Compartor II Operator I	23

D. Stucki

Fight 238-7 STATE (DEPAR IMENT OF I WELL DEPAR IMENT OF I State law requires that this report to filled with	DF IDAHO USE TYPEWRITER OR WATER RESOURCES BALLPOINT PEN ER'S REPORT the Direct Department of Water Semantic
State law requires that this report be filed with within 30 days after the complete	h the Director, Department of Water Resources etion or abandonment of the well.
1. WELL OWNER Name <u>Styry</u> <u>Stycki</u> Address <u>12471 N</u> <u>55 E IDAHU</u> <u>83901</u> Owner's Permit No. 25 - 90 - 15 - 025 - 000	7. WATER LEVEL Static water level <u>fer</u> feet below land surface. Flowing? Yes XB_No G.P.M. flow Artesian closed-in pressure
2. NATURE OF WORK A New well Despened Replacement Well diameter increase Abandoned (describe abandonment procedures such as materials, plug depths, etc. in lithologic log)	Describe artisian or temperature zones below. 8. WELL TEST DATA Pump Bailer Air Other Discharge G.P.M. Pumping Level Hours Pumped
3. PROPOSED USE	
Domestic Irrigation Test Municipal Industrial Stock Waste Disposal or Injection Other	9. LITHOLOGIC LOG Bore Depth Vaterial Vaterial Value
	6' 0 / Clay
A: Rotary Air Hydraulic Reverse rotary Cable Dug Other	1' C Said Clay Cener/ ~ 80' ML' Large Centry + Such X NU' ISO Tighty pack Convert + Such X
5. WELL CONSTRUCTION Casing schedule: [2] Steel Concrete Other Thickness Diameter G	10. Work started 4.4.90 finished 4-6-90
6. LOCATION OF WELL Sketch map location <u>must</u> agree with written location. N Subdivision Name Lot No. Block No. Block No. Scounty <u>Bonne ville</u> JE x <u>NE</u> x Sec. <u>ID</u> , T. <u>3</u> ST R. <u>F</u> w	11. DRILLERS CERTIFICATION I/We certify that all minimum well construction standards were compiled with at the time the rig was removed. Firm Name <u>Accessing Life colling Stree</u> Firm No. 10 Address <u>Birgs 460 (Loor Ed.</u> Date <u>4-6-90</u> Signed by (Firm Official) <u>Jaccessing Electron</u> and (Operator) <u>Jaccessing</u>

J. Stucki

4 18/01		0.0	008118	0												
Delling	Dormit	89	2733							12. 51	ATIC V	VATER	LEVEL and WELL TESTS:		Net.	
Drilling	Permit r	00,	MP	29 MV	V1					Depth	first wat	er enco	untered (ft) 275 Static wa	ater level (ft) _2	275'	_
2 OWN	gni or inj ED.	ection w	ell #							Water	temp. (°	F)_C01	d Bottom hole temp	0. (⁰ F)		_
2. OWN	Idaho	Depa	tment	of Wat	er Res	ource	8			Descri	be acces	ss port_	Locking Vault			_
Name_	. 322	E. Fro	ont St.	o. riai		00100				Well to	ost:	I Die	Tes	t method:	1	2
City B	oise			04	ID		. 8	3720		Draw	down (feel) yi	eld (gom) (minutes) Pur	np Bailer /	Air a	a
2 WELL	1004	TION		31	ale		cip -								4	1
Tun 8	No		or Pour	1. D	n. 1	9 -				Water	quality t	est or ce	omments: 0 PPM Nitrate		-	
Twp	2 140	nin 🗖	or Sou	NI NI	Rge.	SW	ast 🔀	1 or	west 🛄	13. LIT	HOLOG	IC LOC	and/or renairs or abandonn	ant.		-
Sec.			10 60165	-1/4 -10	1/4 Incres	TRD HC	1/4	4		Bore	From	To	Remarks, lithology or description	of regains or	W	Val
Gov't Lo	t	0	County J	erome						(In)	(11)	(ft)	abandonment, water to	mp.	Y	T
Lat.		a	42.741	646	-	(De	g. and D	- Decimal r	inutes)	12	0	19	TOPSOIL			1
Long.		0	114.18	4132		(De:	g, and D	Decimal m	inutes)		19	6/	GRAY BASALT			1
Address	of Well	Site Ca	anal Ro	1							07	10	LOOSE	ST URC.,		4
		6419		Ci	ty Ede	n					78	98	HARD ROCK		-	+
(Give al teast)	Leevie ernen	+ Distance to	Read or Land	rank)							98	118	FRACTURED. SOFT			+
4 1105	B)	n	_ oup. r	vame							118	172	HARD ROCK			t
Dom	estic Г	Munici	Dal XI	Aonitor		tion	Ther	mal F	1 Injection		172	220	FRACTURED GRAY BAS	SALT		t
Other			-						- ingestion	6	220	245	FRACTURED GRAY BA	SALT		1
5. TYPE	OFW	ORK:		331 <u></u>							245	249	MED/SOFT, LOOSE, LO	ST CIRC.		1
X New	well	Replan	cement we	ell 🗆 l	Modify e:	disting w	eli				249	340	FRACTURED ROCK		X	4
6. DRIL	L METH	IOD:														t
X Air R	otary	Mud	Rotary	Cabl	e 🗆 (Other										1
7. SEAL	ING PF	ROCED	URES:								_					1
Bor	material	From	(III) To (II		00 lbe	Plac	ement r	melihod/pi	erubecon							4
Der	Itollite	-		5 5,7	ou ibs	POUR	eu									+
0.040			_			1										+
Diameter	From (th)	To /m	Gauge/	1.40	lerial	Carina	Lines	Threada	d Manual						-	t
(nominal)	+2	220	Schedule 250	Stool	inder sea								RECET	VED		1
	12	220	.200	Steel				-						0000		1
							-	H					MAR 0 3	2020		4
							-	-					DEDT OF WATER S	SOURCES		+
L							Ц	Ц	Ц			-	SOUTHERN H	GION	-	t
Was driv	e shoe	used?	NY D	N Shoe	Depth(s)	220		_								1
9. PERF	ORATI	ONS/S	CREENS	B:												T
Perforati	ions 🔲	YDN	Metho	d								-				1
Manufac	tured so	reen C		Type												4
Method	of install	ation														+
From (ft)	To (ft)	Slot size	Number	n Diamet	er M	laterial	G	auge or s	Schedule	0						1
		1	1	Linemin			1			Compl	eted Dep	m (Meas	surable): 010	4 14 4 10 4 5 5		-
		1	1	-			-			Date S	tarted: 1	/14/20	Date Completer	1/16/2020	2	_
		-		-			-			14. D	RILLER	S CER	TIFICATION:			
Length o	f Heada	ion	1	1	ath of T	ailning	1.			the tin	the the rig	was re	moved.	s were complie	d with	1 8
Packer		IN THE		Ler	igni or 1	ashibe"				Come	ony No-	Elsi	ng Drilling & Pump Co. Inc.	0. H 66	9	
10 61 7		an typ								oump	may reall	1	_6	_ Co. No00		-
Fille	Malerial		am (m)	Torre	Quentite at	is or all	-		mailhort 1	*Princ	pal Drille	er Ly	15-	_ Date1/24/	2020	1
- die	- meyelidi	-10		- 4 ml	uruaniity (il	s or it')	PI	acement	Domen	*Drille	·	Ci		Date 1/24	/2020)
										*00000	ator II			Detr		1
44 51 0	ALL DO	ADTEC								open	a.or 11	n	1101	_ Date	maar	-
11 ELO	WING	ARTES	AN:	Jaganas						Opera	tor I 🟒	-	NU	Date_1/24/	2020	0
	Artaeion	? 🗆 Y	X N A	rtesian P	ressure	(PSIG)				* Sign	ature of	Princip	pal Driller and rig operator are	required		
Flowing	rite alein															

Appendix B – Basalt Site Well Logs

WELL TAC NO D 0081181							
NUMERIC TAG NO. D 0001101	12. ST	ATIC W	ATER	LEVEL and WELL TESTS:			
Water sicht er isissission um # MP29 MW2	Depth	first wate	r encou	intered (ft) 252' Static w	ater level (ft)	245'	
2 OWNER-	Water	temp. (°F	Cold	Bottom hole tem	ρ. (^a F)		_
Name Idaho Dept. of Water Resources	Descri	be access	s port_	Locking vauit			
Addrage 322 E Front St.	Well to	ost:	Dis	Charge or Test duration	st method:		David
City Boise State ID Zin 83720	Draw	down (feet)	yie	ald (gpm) (minutes) Pu	mp Bailer	Air	rtesa
3 WELL LOCATION:			-		4 8 .	H	H
Two 8 North C or South R Rae 19 East R or West	Water	quality te	st or co	mments: 2 PPM Nitrate		-	Ч
Sec. 10 1/4 SE 1/4 SW 1/4	13. LIT	HOLOGI	C LOG	and/or repairs or abandonr	nent:		
10 acres 100 acres	Bore Dia.	From	To	Remarks, lithology or descriptio	n of repairs or	v	later
Gov't Lot County Jerome	(in) 12	0	7	abandonment, water to	imp.	Y	N
Lat 0 42.739116 (Deg. and Decimal minutes)	12	7	104	GRAY BASALT	_	-	
Long. 0 114.221493 (Deg. and Decimal minutes)		104	110	CINDERS		-	1 x
Address of Well Site South Eden Ru		110	123	BROWN ASH & CINDER	RS		X
Give at treast name of read + Distance to Read or Landmarks		123	170	GRAY BASALT			X
Lot Blk Sub. Name		170	194	FRACTURED GRAY BA	SALT		X
4. USE:		194	200	GRAV BASALT	RC.	-	X
Domestic Municipal Monitor Infigation Thermal Injection	6	200	237	GRAY BASALT		-	H
5. TYPE OF WORK	_	237	240	FRACTURED GRAY BA	SALT	-	1×
New well Replacement well Mcdify existing well		240	252	GRAY BASALT			X
Abandonment Other		252	261	FRACTURED BROWN E	BASALT &		
		261	300	CINDERS	CALT	X	
		201	500	RACTORED GRAT DA	SALT	-	+-
Seal material From (ft) To (ft) Quantity (bs or ft ²) Placement method/procedure						-	-
Bentonite 0 200 8,150 lbs Poured							
8. CASING/LINER:						-	
(nominal) From (ft) To (ft) Schedule Material Casing Liner Threaded Welded						-	+
6 +2 200 .250 Steel	-			DECEL	FD	-	\vdash
				MEUEIN			
				110 0 2 2	020		
				MARUJZ	520	-	-
Was drive shoe used? X Y IN Shoe Depth(s) 200'	-			DEPT OF WATER DE	SOURCES	-	⊢
9. PERFORATIONS/SCREENS:			_	SOUTHERN HE	GIUN	-	+
Perforations Y N Method							
Manufactured screen Y N Type							
Method of installation						-	-
From (ft) To (ft) Stot size Number/ft Diameter Material Gauge or Schedule				300'		_	_
(nominal)	Comple	eted Depth	n (Meas	urable): 000	1/20/202	-	
	Date S	tarted: 1/	21/20	Date Complete	d: 1/28/202	0	_
	14. Di	RILLER'S	S CER	TIFICATION: imum well construction standard	to woro come	المراجع المراجع	
Length of Headpipe Length of Tailpipe	the tim	e the rig	was ren	noved.	is were comp	ieu with	at
Packer Y N Type	Comp	any Name	Elsir	ng Drilling & Pump Co Inc	Co No 6	69	
10.FILTER PACK:	*Dringi	ingl Driller	C	2	1/3	0/2020	_
Filter Material From (ft) To (ft) Quantity (los or ft ²) Placement method	Philip	pai Uniter	1	8	_ Date	0/2020	_
	*Drille	·	0		_ Date	0/2020	_
	*Opera	ator Ii			Date		
11. FLOWING ARTESIAN:	Opera	tor	n	>AII/	Data 1/30	0/2020	
Flowing Artesian?		-	.1	W			_
		stress of I	and shall be	al Deilles and de anostation and			

Form 238-7 11/97 JGE To Concern and the second	REPORT CITIC Use Only	
WELL TAG NO. D 0000955 DRILLING PERMIT NO. Other IDWR No. 764880	REFURI / Two Page San	
DRILLING PERMIT NO		
Other IDVVR 140. 704000 MVV1	11. WELL TESTS:	:
2 OWAIED.	Pump LiBaller XJAir LiFkowing Artesian Yieki gal/min. Drawdown Pumping Level Tim	ne
Vame Idaho Dept. of Water Resources	no returns	
City Boise State [D Zip 83706		
3. LOCATION OF WELL by legal description:	Water Temp. <85 Bottom hole temp	
Sketch map location must agree with written location.	Depth first Water Encounter	- ,
	Bore	w
Ree. 18 East X or West	Dia, From To Remarks: Lithology, Water Quality & Temperature	+ Y
Sec. 26 1/4 SE 1/4 SW 1/4	10 2 3 black lava	_
Gov't Lot County Jerome	10 3 4 brown clay	+
S Address of Well Site	8 19 114 gray lava	+
City Jerome	8 114 broken , black lava & red	1.
(Give at least nume of road + Olitizate to Road or Landmark) t. Bilk. Sub Name	8 119 medium hard lava &	+
	124 red/black cinders	\pm
I USE:	8 124 179 black lava	-
Themal Injection Other	8 179 188 medium hard lava & red ash	+
	8 191 195 soft & broken (no returns)	╈
XINew Well Modify Abandonment Other (Replacement etc.)	8 195 253 hard lava	
	8 253 256 cinders	+-
S. DRILL METHOD:	8 272 276 cinders	+
	8 276 281 hard lava	
SEALING PROCEDURES:	8 281 284 hard & broken lava	X
Seal/Filter Pack AMOUNT METHOD	8 301 304 soft & broken lava	Tx
hentonite acoulo C 0 19 2001bs dry pour	8 304 311 hard lava	
bentonite 38 Aug 0, 262, 99 bass dry pour	8 311 312 broken lava	+
Pea grave 262 310 1-5 yrds. dry pour		+-
Vas drive shoe used?		1.
Vas drive shoe seal tested? []Y XN How?		+
. CASING/LINER:	DECEIVED	1
Stameter From To Guage Material Casing Liner Welded Threaded	RECEIVED RECEIVED	+
4" 0 270 PVC 🖾 🗆 🖾	OCT - 3 2000	+
	ULI 3 ZU00	_
ength of Headpipe 1' Length of Tailpipe Con	artment of Water Hesources Southern Repton	+
PERFORATIONS/SCREENS:		+
Screens Screen Type		
From To Sint Size Number Diameter Material Coolea Liner	Completed Depth 310* (Mean Date: Stated 5/27/2000	isural
270 310 102 4 ^j PVC X	541. 01.120 <u>672672000</u> 00.1172000	
	 DRILLER'S CERTIFICATION: IWe certify that all minimum well construction standards were complied with the time the rig was removed. 	h at
	Company Name Eaton Drilling & pump Firm No. 26	<u>}</u>
D. STATIC WATER LEVEL OR ARTESIAN PRESSURE:	Firm Official	2000
epth flow encountered ft. Describe access port or control	and and and a fille	U
evices: well cap	Driller or Operator	2000

East Well

IDAHO DEPARTMENT OF WATER RESOURCES WELL DRILLER'S REPORT

1. WELL TAG NO. D 0070631	12. ST	TATIC W	ATER	LEVEL and WELL TESTS:		
Drilling Permit No. 880463	Depth first water encountered (ft) 287' Static water level (ft) 277'					
Water right or injection well #	Water temp. (°F) Bottom hole temp. (°F)					
2. OWNER:	Descri	be acces	s port_			
Name Idaho Dept.of Water Resources	Well te	est:		Test method:		
Address 322 E. Front Street	Draws	dawn (feet)	Dis	charge or Test duration Pump Bailer Ai	ir F	lowing
City Boise State Idaho Zip 83720-0098	No F	Return			ব্ৰ ঁ	
3.WELL LOCATION:			ł			
Twp. 8 North 🔲 or South 🗵 Rge. 19 East 🗵 or West 🗋	Water	quality te	st or co	omments:		
Sec. 2 1/4 NW 1/4 SW 1/4	13. LIT	HOLOG	IC LOG	and/or repairs or abandonment:		-
10 m2 ms 40 m2 ms - 160 m2 ms	Dia.	From (ft)	To (ft)	Remarks, lithology or description of repairs or abandonment, water temp.	W:	Iter
Gov'i Lot County Jerome	12"	0'	42'	Grav Basalt		v
Lat. 42 0 45.312 (Deg. and Decimal minutes)	12"	42"	50'	Black Basalt Cracked		x
Long, 114 012.521 (Deg. and Decimal minutes)	12"	50'	118'	Black Basalt Caving Lost Circ.		X
Address of Well Site	12"	118'	62'	Black Basalt Circ. Returned		х
City Eden	_12"	62'	118'	Black Basalt Hard		х
Lot Blk Sub. Name	12"	118'	130'	Black Basalt Soft Caving Lost Circ.		x
4. USE:	12"	130	165	Basalt Hard		X
Domestic Municipal Monitor Irrigation Thermal Injection	12"	175'	205'	Basalt Loose & Caving		X
X Other Observation	12"	205'	230'	Basalt Hard		X
5. TYPE OF WORK:	12*	230'	233'	Basalt Soft		x
Abandonment Other	12"	233'	239'	Basalt Hard		x
6. DRILL METHOD:	6"	239'	248'	Basalt Hard		х
X Air Rotary I Mud Rotary Cable Other	6"	248'	254'	Basalt Soft		×
7. SEALING PROCEDURES:	6"	254'	287'	Basalt Hard		x
Seal material From (ft) To (ft) Quantity (lbs or ft ²) Placement method/procedure	6"	287'	291'	Basalt Soft Cracked	X	
3/8 Chip Bent. 3 237 12500 lbs Dry Pour	6"	291	325	Basalt Hard		X
	0	323	340	Circ Returned for 10' then Lost again		
8. CASING/LINER:				Circ. Retained for to their Lost again	<u>^</u>	
(nominal) From (ft) To (ft) Schedule Material Casing Liner Threaded Welded						
6" +1 239' .250 Steel						
				DEALWER		
		-		RECEIVED		
Was drive shoe used? X X I N Shoe Depth/s) 239'				007 4 0 0000		
9 PERFORATIONS/SCREENS:				UCI 1 9 2010	_	
				DEPT OF WATER RESOURCES		-
				SOUTHERN REGION		
Manufactureb screen T T N Type						
Method of Installation						
From (It) To (It) Slot size Number/It (nominal) Material Gauge or Schedule	Comple	ted Deptr	n (Measu	urable):340'		
	Date St	oc hetra	:t. 05, 1	2016 Date Completed Oct. 13, 201	6	
	14. DR	ULLER'S	CERT	TEICATION		
	I/We ce	ertify that	all mini	mum well construction standards were complied	with a	at
Length of Headpipe Length of Tailpipe	the time	e the rig v	was rem	noved.		
Packer Y N Type	Compa	iny Name	A&B	Irrigation District Co. No. 719		
10.FILTER PACK:	*Drincir	al Drillor	ix	Sill V CMANER Date Oct. 14	. 201	6
Filter Material From (ft) To (ft) Quantity (lbs or ft ³) Placoment method	Princip	Jai Driver	<u>~~</u>	Date Control Date		_
	*Driller			Date		_
	*Opera	tor II		Date	_	_
11 ELOWING ARTESIAN:	0	7	1- 1	B Oct 14	1. 201	6
Figure Adapted AV 201 Adapte Proving (0010)	Operation			Date Oct. 14	.,	-
Proving Artesian () Y IXIN Artesian Prossure (PSIG)	* Signa	ture of F	rincipa	al Driller and rig operator are required.		
Describe control device						

Central Well

Frim 23.21 Apple 2 LiDAHO DEPARTMENT OF WATER RESOURCES International and the second of the sec	Dawn 54 202 - 202139	Leastion Corrected by I	
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fit. Describe access port or control and Driller or Operator (Sign once if Fith Official & Operator) FORWARD WHITE COPY TO WATER RESOURCES	S. DRILL METHOD: Mar Rotary Cable Mud Rotary Other 7. SEALING PROCEDURES: Seal/Filter Pack AMOUNT METHOD Material From To Seades or Material From To Seades or Dentonite Amount METHOD Bentonite From To Seades or Dentonite Arconulgr 1 274 325 Cry Pour Deft vie shoe used? Y Xin Shoe Depth(s) Material Method Vas drive shoe used? Y Xin How? Shoe Depth(s) Material Casing Liner Weided Threaded 3. CASING/LINER: Bentonite Casing Liner Xin Image: State of the st	6 186 209 Diack lava 8 209 231 medium hard black lava 8 231 238 black cinders 8 231 238 black lava 8 231 238 black lava 8 231 238 black lava 8 249 black lava indering the state of the	X X X X X X X X X X X X X X X X X X X
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FORWARD WHITE COPY TO WATER RESOURCES	6. DRILL METHOD: XAir Rotary Cable Mud Rotary Other 7. SEALING PROCEDURES: Seal/Filter Pack AMOUNT METHOD bentonite from To Stade of Boundard Method bentonite arcouldr 0 19 200 lbs. dry pour bentonite arcouldr 2.74 325 1 yr.o. dry pour Nas drive shoe.used? Y Xin Shoe Depth(s) Material Casing Iner As drive shoe seal tested? Y Xin How? Iner Xin Social State Kin Barneter From To Guage Material Casing Iner Xin 4.* 0 284 PVC Xin Xin Xin Xin Xin	6 186 209 Diack lava 8 209 231 medium hard black lava 8 231 238 black cinders 8 231 252 medium hard royolite & 301 medium hard red/black lava 308 & breaks 8 301 medium hard red/black lava 308 8 308 & breaks 8 8 308 & breaks 8 8 308 325 hard gray ryolite 9 0C1 - 3 2000 0C1 - 3 9 0 0 Completed B/27/2000 13 DRILLE	x x x x x x x x x x x x x x x x x x x
	B. DRILL METHOD: Mul Rotary Coher Other Other ScallFilter Pack AMOUNT METHOD ScallFilter Pack AMOUNT METHOD Material From To Support Dentonite acrowler 0 19 200 lbs. dry pour bentonite acrowler 0 19 200 lbs. dry pour Dentonite acrowler 0 19 200 lbs. dry pour Dentonite acrowler 0 19 200 lbs. dry pour Dentonite acrowler 0 19 200 lbs. dry pour Dentonite acrowler 0 19 200 lbs. dry pour Dentonite acrowler 0 19 200 lbs. dry pour Dentonite acrowler 0 19 200 lbs. dry pour Dentonite acrowler 0 19 200 lbs. dry pour Dentonite acrowler 0 19 20 lbs. dry pour dry pour Dentonite acrowler 0 19 20 lbs. dry pour dry	6 186 209 Diack lava 8 209 231 medium hard black lava 8 231 238 black cinders 8 231 238 black cinders 8 231 238 black cinders 8 231 238 black lava 8 231 252 redium hard ryolite & 8 252 medium hard red/black lava 8 301 medium hard red/black lava 308 & breaks 8 8 308 325 hard gray ryolite 9 0C1 - 3 2000 0C1 - 3 2000 0C1 - 3 0C1 - 5 -2000 Department of Water Resources 0 0C1 - 3 2000 0CT 5 - 2000 Department of Water Resources 0 0C1 - 3 2000 0CT 5 - 2000 Department of Water Resources 0 0.25/2000 Completed 8/27/2000 13. DRILLER'S CERTIFICATION: We certify that all minimum well construction standards were complied withe the time the rig was removed. Company Name Eatop Drill	x x x x x x x x x x x x x x x x x x x

West Well (corrected from MW2 as listed)

Appendix C – 07N 39E 07 BDA1 Well Log

A-2

THE OF NEVE	ALLON - REGION				SHEET 1 OF 1
			LOG O	F W	ELL Fremont Co.
Project Teton Ba	sin Project	Feature	Observation	n Hole (Egin Lakes Pilot Recharge Project)
Wall his 78/205	- 7 bds 1 (C 1)			17021 0	State Idano
Well 140. 141 320	- 7 044 1 (0-1)	Locat	ion Approx.	1793. 2	. 2370' E, NN corner Sec. 7, T7N, R39E
Total Depth 3	40	Begun 7/24/	73 Com	_betslqn	8/15/73 Drilling Method Cable Tool
Static Water Le	evel_58.46		NELWOOK Mer	os. Pt.	Top 6" casing Date 4/26/74
Elevation (ground	4874.5		W L Mooo	De	A4(ghove)
Vield	D /		W. L. Meus.	. Pl	Feet 2000000 Groun
11010	Orowdown_		Other	Doto_	Driller's Reports
Logged By Has	kettG	sophysical L	NoneNone		Drilled By Cushman & Denning
Dritting Data	Description	Well	14	a 2 9 0	
Water' Samples	of Well Completion	Diagram	1 Log	Type Con	Classification and Physical Condition
Spec. No. 100C-12	68	631	1 1. 1. 1. 1.		
	Set temporary	10" hole	115		0-5' Sand, gray
or111 Rig 22-W	10" csg. to 5'.	20' 4	111	· · ·	10-20' with brown streaks
		ar sala F	1.15		42-45' Basalt, red, broken with sand wad
		++ +5 -	19-51		45-54' Sand, red with small gravel
water levels	- ·				54-64' Basalt, gray
7/30/73 53.5'		6" hole	+= 573	-	68-163' Bacalt consu
nole at 68', 6"	caving, cemented	Acom and m	C.C.C.C.		82-110' porous, with cinders
Lay. At 03'.	back to 70'	dail 8 to 1	80		110-125' porous
7/31/73 53.5'	grouting 6" csg.		1 1 1 +		*
8/1/73 53.5'			1:+5		
hole at 75'. 6"		6" hole (LE.M.		
sg. at 75'.	A+ 3401 hadlad	depth)	120 ->		
8/4/73 53.5'	2 hrs. to clean	(340'). (1 1		163-183' Basalt, red and gray
hole at 125', 6"	and develop.	1	13-1-		192 1001 Class Laws
csg. at 75'.			1		103-100 Clay, Drown
8/7/73 54.0'		(1601 5		188-340° Basalt, gray
hole at 125', 6"	Backfilled	(1 the		213-218' porous, with cinders
csg. at /5'.	with cuttings	. 1	TTE		238-260' porous, with cinders
:	70-18";	(The set		260-300' porous
4/25/74 58,60	to surface.	1	200 2.		300-340' porous, with cinders
4/26/74 58,46	pulling 10" csg.		3-5.		-
			4 · *		
		1	15.1		
1		{	240 :		
		. 1	1:22		340' Total Depth
		1	1		
1		1	1 10:5		Driller's Notes:
		. 1	280 . (*		42-54' caving, reamed and drilled to 75' and set 6" casing at 75'.
		(1-25		82-110' Cuttings washing into formation
12.	-incha)	12.0		300-340' appears to be a lot of water
M.P 4875.94	rocke	cap	1. * 2		werears to be a 101.01 Water
4875.27-	-top	6" csg	220 × . ["		
. 4874.57	E E	top, slai	Par(.	
£	they gains	Ground	0.7020		
·		sfc.	W 111 W		-
sur	Face A B		1005		* XI
	seal 6" c	sing	1		
25	. 4 9				
	1 (1)				
		•			×
SAMPLE TYPE:		WTI a.	1		
CR = Core CT = Cuttings	DAS	Sand	22	Clay	Basalt, vesicular
D Drillers Log		Gravel	[2	ST. Basa	It Esta Cinders

ì.
M723923	
WELL TAG NO. D (2) (3)	12. STATIC WATER LEVEL and WELL TESTS:
Drilling Permit No.	Depth first water encountered (ft) Static water level (ft)
Water right or injection well #	- Water temp. (°F) Bottom hole temp. (°F)
Now Bar CL Manager t	Describe access port
Address 1405 Itallinge Dr.	Well test: Test method:
city Tacho Falls State ID Zip 8340/	- Chandown (red) - yield (gom) (minutes) - unity baller All artesia
WELL LOCATION:	
wp. 🔽 North 🛛 or South 🗖 Rge. <u>38</u> East 🖾 or West 🗖	Water quality test or comments:
ec. 12 1/4 Sh 1/4 SE 1/4	13. LITHOLOGIC LOG and/or repairs or abandonment:
	Dia. From To Remarks, lithology or description of repairs or Water (in) (ft) (ft) abandonment, water temp. Y
Sovi Lot County Frence (10° 0' 4' Querburden
at	4' 35' Basalt
ddress of Well Site 1.5 Mikou 2 state Ground Station	55 40' Soft, Red sediment
Egin Hamer Rod. City	85 89' Soft black basalt
ot Blik Sub Name	89' 105 Med based & water
	105 113' Soft besalt more water
Domestic 🗋 Municipal 🙀 Monitor 🗋 Irrigation 🗋 Thermal 🔲 Injectio	113 128 Mad basalt
Other	132/13/ Soft baselt, Seducet Cours
. TYPE OF WORK:	142 163 Med. Ensult
Abandonment Other	153 154' Soft
DRILL METHOD:	154 158 Med. Ensait
Air Rotary Mud Rotary Cable Other	196 195 Course Cinders red ; black layers
Seal material [From (ft)] To (ft) [Quantity (lbs or ft ²)] Placement method/procedure	202'207 Course Cinders
Sentonte O' 165' 8500# Quebore	207' 215' Med Easait
	215 232 Caving Cinders
. CASING/LINER:	232 235 Med basalt
nominal) From (ft) To (ft) Schedule Material Casing Liner Threaded Welded	240'270' Soft basalt
6" +2' 238.25 steel 14 11 14	
/as drive shoe used? 🔁 Y 🔲 N Shoe Depth(s)	BECEIVED
PERFORATIONS/SCREENS:	
erforations I Y I N Method	UEC 14 /0/0
anufactured screen Y N Type	Evapartment of Water Resources
lethod of installation	Fastesu naĝion
From (ft) To (ft) Slot size Number/ft Diameter (nominal) Material Gauge or Schedule	Completed Depth (Measurable): 270
	Date Started: 11/20 Date Completed: 11/20/20
	14. DRILLER'S CERTIFICATION:
	I/We certify that all minimum well construction standards were complied with at the time the rig was removed.
ength of Headpipe Length of Tailpipe	and Dellas an SIG
	Company Name Co. No. JI
Ether Material Error (#) To (#) Countils (the scale)	*Principal Driller och Ennine Date
Placement method	Driller Led 1 Lang Date 1/30/20
	*Operator II Date
	Operator I Data
LI LOTHING ARTEGIAN.	Operator 1 Date

Appendix D – Combination Alluvial/Basalt Site Deep Well Logs

Easement Well Deep

1. WELL TAG NO. D CO 81455	12 STATIC WATER LEVEL and WELL TERTE.
Drilling Permit No.	Depth first uniter encountered (#) /// 2 Static uniter level (#) /// 2
Water right or injection well #	Water temp. (^o E) Bottom hole temp. (^o E)
2. OWNER:	Describe access port
Name Idaho Water Resource Board	Well test: Test method:
Address B Bux 83720	Drawdown (feet) Discharge or Test duration Pump Bailer Air Flo
CityStateZip	- St 30 0 0 &
3.WELL LOCATION:	
Twp North Q or South Rge. S East & or West	Water quality test or comments:
Sec	Bore From To Remarks, lithology or description of repairs or Wa
Sov't Lot County Fremont	(in) (ft) (ft) abandonment, water temp. Y
at. 43" 56,526 (Deg. and Decimal minutes)	10 0 1 Sand
ong (Deg and Deginal minutes)	54'55' Contraction
Address of Well Site 225 miles West of theend of	56 95' Soft Lesalt / Very Endund
the presencent on Egin Howevicity Road	95' 100' Soft basalt, uniform
.ot Blk Sub. Name	100 153 Medium pacalt, uniform
4. USE:	193 157 Broken basalt
Domestic Municipal Monitor Irrigation Thermal Injection	177' ZIL' Hand Lasal - Gradie Drilling
TYPE OF WORK:	216 228 Pluged a little at 216
New well Replacement well Modify existing well	Soft Spreken w/clay
Abandonment Other	228 237 Have bagalt
Air Rotary 🔲 Mud Rotary 🔲 Cable 🔲 Other	251 275 Enclund bacalt
SEALING PROCEDURES:	
Seal material From (ff) To (ff) Quantity (bs or ff?) Placement method/procedure	
Sutonile O' 253 7900# aubore	
Diameter From (11) To (11) Gauge/ Material Casino Liner Threaded Welder	
6 +7' 253 25 st-0 KB 0 0 B	
	DECENTER
Vas drive shoe used? K X U N Shoe Death(a)	neverved
	AUG 8 3 2828
	Exactment of Water Permises
	Lastern Region
tethod of installation	
Diamater	
rom (N) To (N) Slot size Number/II (nominal) Material Gauge or Schedule	Completed Depth (Measurable): 275 '
	Date Started: 6 29/20 Date Completed: 7/2/20
	14. DRILLER'S CERTIFICATION:
	I/We certify that all minimum well construction standards were complied with a the time the rig was removed.
ength of Headpipe Length of Tailpipe	Commentant and the and star
	Company Nance and I sharing Co. No.
Filter Material From (8) To (8) Complify (the or 6%) Blockmont with the	*Principal Driller
Promiting To (ii) Guantity (ibs of it.) Placement method	Diller Date 7/7/20
	*Operator II Date
	Unit Date
	Operator I Date

West Well Deep

Г

Appendix E Weekly SC							
Measurements							
Sample Location	Date	Electrical conductivity					
		uS/cm					
	9/12/2019	253					
Headgate	9/20/2019	229					
at the	9/27/2019	299					
alluvial	10/3/2019	290					
site	10/11/2019	309					
The	10/17/2019	345					
Jones Pit	10/25/2019	347					
	10/31/2019	367					

Appendix E – Additional Headgate SC values from the alluvial site – The Jones Pit

Appendix F – Assessment of uncertainties for various constituents, endmember mixing fractions (f_1) and $\sum Cation_{excess}$

The relative uncertainty in the concentrations of groundwater constituents was estimated from seven samplings of the West and MW2 wells (Table 5) at MP29/31 located 5,930 and 6,775 ft (1,810 and 2,065 m) from the basin over a period of 227 days. The composition of water samples from both these wells appeared to be minimally influenced by MAR activities. The relative standard deviations for each constituent as well as that for the sodium adsorption ratio (SAR) and the equivalence base sulfate to chloride ratio calculated for each well are given in Table F.1. The average of the two relative standard deviations for each constituent or ratio was used as an estimate of the relative uncertainties associated with sampling, analytical methods, and spatial variations at a given location (i.e., different locations may have different concentrations on the same date).

Table F.1 -- The relative standard deviations among different constituents and ratios from the West Well and MW2 at the basalt site – MP29/31, and their average which was used to estimate the relative uncertainties associated with repeated samplings.

Constituent/Ratio	Relati	ve Standard Deviation		
	West	MW2	Average	
Ca ²⁺	2.8%	2.6%	2.7%	
Mg ²⁺	3.0%	4.0%	3.5%	
Na⁺	2.4%	2.1%	2.2%	
K⁺	1.6%	1.1%	1.4%	
Cl ⁻	2.6%	2.0%	2.3%	
SO4 ²⁻	2.6%	1.9%	2.3%	
Alkalinity	2.4%	7.5%	5.0%	
SAR	2.5%	2.0%	2.2%	
SO4 ²⁻ /Cl ⁻	2.4%	2.4%	2.4%	

Uncertainties associated with mixing calculations were estimated from a simulated normally distributed random population of 10,200 samples calculated using Microsoft Excel. The simulated sample population had the same mean concentrations as the Scenario 4 (mixing, cation exchange, and calcite precipitation) Jones Pit GWB modeling results for mixture of 50% ambient groundwater and 50% infiltrated surface water and relative standard deviations given in Table F.1The mixing fractions and [species]_{excess} values were then calculated for each of the simulated samples the absolute uncertainties (Table F.2) estimated from the population standard deviations are given in Table F.2.

	Absolute Uncertainty (±1 St. Dev.)	Units
f_1	0.12	None
Na⁺	0.04	meq/L
Ca ²⁺	0.09	meq/L
Mg ²⁺	0.09	meq/L
Alkalinity	0.10	meq/L
∑Cation _{excess}	0.12	meq/L

Table F.2 – Uncertainties for various constituents, endmember mixing fractions (f_1) and $\sum Cation_{excess}$.

Although the estimated uncertainties are specific to Jones Pit water composition, they are not particularly sensitive to the actual water composition as long as the difference between the concentration of the conserved ion used in the mixing calculations. As such these estimated uncertainties are applicable for all three sites.

Scenario 1	Mixing							
Carbonate alkalinity	Ca++	CI-	HCO3-	K+	Mg++	NO3-	Na+	SO4
meq_acid/l	mmol/l							
3.6823246	1.584	0.2331	3.847656	0.05627	0.576	0.02278	0.4143	0.42995
3.6553461	1.575434	0.243031	3.812825	0.056266	0.575956	0.021875	0.422041	0.435069
3.6294205	1.567212	0.252564	3.779388	0.056262	0.575913	0.021005	0.429472	0.439983
3.6045053	1.559318	0.261717	3.747288	0.056258	0.575872	0.020171	0.436607	0.444701
3.5805594	1.551739	0.270503	3.716472	0.056254	0.575833	0.01937	0.443455	0.44923
3.5575438	1.544464	0.278937	3.686889	0.05625	0.575795	0.018601	0.45003	0.453577
3.5354207	1.53748	0.287035	3.658489	0.056246	0.575759	0.017863	0.456342	0.457751
3.5141542	1.530775	0.294808	3.631225	0.056243	0.575725	0.017154	0.462402	0.461758
3.4937098	1.524339	0.302271	3.605051	0.05624	0.575691	0.016474	0.468219	0.465605
3.4740545	1.518159	0.309435	3.579925	0.056237	0.575659	0.015821	0.473803	0.469297
3.4551565	1.512227	0.316312	3.555803	0.056234	0.575629	0.015193	0.479164	0.472843
3.4369857	1.506533	0.322914	3.532646	0.056231	0.575599	0.014592	0.48431	0.476246
3.419513	1.501066	0.329253	3.510416	0.056228	0.575571	0.014014	0.489251	0.479513
3.4027108	1.495818	0.335337	3.489075	0.056225	0.575544	0.013459	0.493994	0.482649
3.3865525	1.490779	0.341179	3.468587	0.056223	0.575518	0.012926	0.498547	0.48566
3.3710127	1.485942	0.346786	3.448919	0.05622	0.575492	0.012415	0.502919	0.488551
3.3560673	1.481299	0.35217	3.430038	0.056218	0.575468	0.011924	0.507115	0.491326
3.3416931	1.476842	0.357338	3.411912	0.056216	0.575445	0.011453	0.511143	0.49399
3.3278678	1.472562	0.362299	3.39451	0.056214	0.575423	0.011001	0.515011	0.496547
3.3145702	1.468454	0.367062	3.377805	0.056212	0.575402	0.010567	0.518724	0.499002
3.3017802	1.46451	0.371634	3.361769	0.05621	0.575382	0.01015	0.522288	0.501359
3.2894781	1.460724	0.376024	3.346373	0.056208	0.575362	0.009749	0.525709	0.503622
3.2776456	1.45709	0.380238	3.331594	0.056206	0.575343	0.009365	0.528994	0.505794
3.2662646	1.4536	0.384283	3.317405	0.056204	0.575325	0.008996	0.532147	0.507879
3.2553182	1.450251	0.388167	3.303784	0.056202	0.575308	0.008642	0.535175	0.509881
3.24479	1.447035	0.391895	3.290708	0.056201	0.575291	0.008303	0.538081	0.511802
3.2346643	1.443948	0.395474	3.278155	0.056199	0.575275	0.007976	0.540871	0.513647
3.2249259	1.440985	0.39891	3.266104	0.056198	0.57526	0.007663	0.543549	0.515418
3.2155605	1.438139	0.402208	3.254535	0.056196	0.575245	0.007362	0.54612	0.517119
3.2065541	1.435408	0.405375	3.243429	0.056195	0.575231	0.007074	0.548588	0.518751
3.1978932	1.432786	0.408415	3.232767	0.056194	0.575217	0.006796	0.550958	0.520318
3.1895651	1.430269	0.411333	3.222532	0.056192	0.575204	0.00653	0.553233	0.521822
3.1815572	1.427853	0.414135	3.212706	0.056191	0.575192	0.006275	0.555417	0.523266
3.1738577	1.425533	0.416824	3.203273	0.05619	0.57518	0.00603	0.557513	0.524652
3.1664551	1.423306	0.419406	3.194217	0.056189	0.575168	0.005794	0.559526	0.525983
3.1593383	1.421168	0.421885	3.185524	0.056188	0.575157	0.005568	0.561458	0.527261
3.1524966	1.419116	0.424264	3.177178	0.056187	0.575146	0.005351	0.563312	0.528487

Appendix G – Modeling results for the alluvial site – The Jones Pit for Six Scenarios

3.1459197	1.417145	0.426548	3.169166	0.056186	0.575136	0.005143	0.565093	0.529665
3.1395977	1.415254	0.428741	3.161475	0.056185	0.575127	0.004943	0.566802	0.530795
3.133521	1.413438	0.430847	3.154091	0.056184	0.575117	0.004751	0.568443	0.53188
3.1276804	1.411695	0.432868	3.147003	0.056183	0.575108	0.004567	0.570019	0.532922
3.122067	1.410022	0.434808	3.140198	0.056182	0.575099	0.00439	0.571531	0.533922
3.1166722	1.408415	0.43667	3.133665	0.056181	0.575091	0.00422	0.572983	0.534882
3.1114877	1.406873	0.438458	3.127394	0.05618	0.575083	0.004057	0.574377	0.535804
3.1065056	1.405392	0.440175	3.121374	0.05618	0.575075	0.003901	0.575715	0.536689
3.1017182	1.403971	0.441823	3.115594	0.056179	0.575068	0.003751	0.576999	0.537538
3.097118	1.402606	0.443405	3.110045	0.056178	0.575061	0.003606	0.578233	0.538354
3.092698	1.401296	0.444923	3.104719	0.056178	0.575054	0.003468	0.579416	0.539136
3.0884512	1.400039	0.446381	3.099605	0.056177	0.575048	0.003335	0.580553	0.539888
3.0843709	1.398832	0.447781	3.094696	0.056176	0.575041	0.003207	0.581644	0.540609
3.0804509	1.397673	0.449125	3.089984	0.056176	0.575035	0.003085	0.582691	0.541302
3.076685	1.39656	0.450414	3.08546	0.056175	0.57503	0.002967	0.583697	0.541967
3.0730671	1.395492	0.451653	3.081116	0.056175	0.575024	0.002854	0.584662	0.542605
3.0695917	1.394467	0.452841	3.076947	0.056174	0.575019	0.002746	0.585589	0.543218
3.0662531	1.393482	0.453983	3.072944	0.056174	0.575014	0.002642	0.586478	0.543806
3.0630462	1.392537	0.455078	3.069102	0.056173	0.575009	0.002542	0.587332	0.544371
3.0599657	1.39163	0.45613	3.065413	0.056173	0.575004	0.002446	0.588152	0.544913
3.0570068	1.390759	0.45714	3.061872	0.056172	0.575	0.002354	0.588939	0.545433
3.0541648	1.389923	0.458109	3.058472	0.056172	0.574995	0.002266	0.589695	0.545933
3.0514351	1.389121	0.459039	3.055209	0.056171	0.574991	0.002181	0.59042	0.546413
3.0488133	1.38835	0.459933	3.052075	0.056171	0.574987	0.002099	0.591116	0.546873
3.0462953	1.387611	0.46079	3.049068	0.056171	0.574983	0.002021	0.591785	0.547315
3.0438769	1.386901	0.461614	3.04618	0.05617	0.57498	0.001946	0.592426	0.54774
3.0415542	1.386219	0.462404	3.043408	0.05617	0.574976	0.001874	0.593042	0.548147
3.0393236	1.385564	0.463163	3.040747	0.05617	0.574973	0.001805	0.593634	0.548538
3.0371814	1.384936	0.463891	3.038193	0.056169	0.57497	0.001739	0.594202	0.548914
3.0351241	1.384333	0.46459	3.03574	0.056169	0.574966	0.001675	0.594747	0.549274
3.0331485	1.383754	0.465261	3.033386	0.056169	0.574963	0.001614	0.59527	0.54962
3.0312512	1.383198	0.465906	3.031126	0.056168	0.574961	0.001555	0.595772	0.549952
3.0294293	1.382665	0.466525	3.028956	0.056168	0.574958	0.001498	0.596254	0.550271
3.0276797	1.382152	0.467118	3.026873	0.056168	0.574955	0.001444	0.596717	0.550577
3.0259996	1.381661	0.467689	3.024873	0.056168	0.574953	0.001392	0.597162	0.550871
3.0243863	1.381189	0.468236	3.022953	0.056167	0.57495	0.001342	0.597588	0.551153
3.022837	1.380735	0.468761	3.021111	0.056167	0.574948	0.001295	0.597998	0.551424
3.0213494	1.3803	0.469266	3.019341	0.056167	0.574945	0.001249	0.598391	0.551684
3.019921	1.379883	0.46975	3.017643	0.056167	0.574943	0.001204	0.598769	0.551934
3.0185493	1.379482	0.470215	3.016012	0.056167	0.574941	0.001162	0.599131	0.552173
3.0172323	1.379097	0.470661	3.014447	0.056166	0.574939	0.001121	0.599479	0.552403
3.0159676	1.378727	0.47109	3.012945	0.056166	0.574937	0.001082	0.599813	0.552624
3.0147533	1.378372	0.471501	3.011502	0.056166	0.574936	0.001045	0.600134	0.552836

3.0135874	1.378032	0.471896	3.010117	0.056166	0.574934	0.001009	0.600441	0.55304
3.0124678	1.377705	0.472275	3.008788	0.056166	0.574932	0.000974	0.600737	0.553235
3.0113929	1.377391	0.472639	3.007512	0.056166	0.57493	0.000941	0.60102	0.553423
3.0103608	1.37709	0.472988	3.006286	0.056165	0.574929	0.000909	0.601293	0.553603
3.0093698	1.376801	0.473323	3.00511	0.056165	0.574927	0.000879	0.601554	0.553775
3.0084183	1.376523	0.473645	3.003981	0.056165	0.574926	0.000849	0.601805	0.553941
3.0075048	1.376256	0.473954	3.002897	0.056165	0.574925	0.000821	0.602046	0.554101
3.0066276	1.376	0.474251	3.001856	0.056165	0.574923	0.000794	0.602277	0.554254
3.0057855	1.375755	0.474536	3.000857	0.056165	0.574922	0.000768	0.602499	0.5544
3.0049769	1.375519	0.474809	2.999898	0.056165	0.574921	0.000743	0.602712	0.554541
3.0042005	1.375292	0.475072	2.998978	0.056164	0.57492	0.000719	0.602917	0.554677
3.0034551	1.375075	0.475324	2.998094	0.056164	0.574918	0.000696	0.603113	0.554807
3.0027395	1.374866	0.475566	2.997245	0.056164	0.574917	0.000674	0.603302	0.554931
3.0020524	1.374666	0.475798	2.996431	0.056164	0.574916	0.000653	0.603483	0.555051
3.0013927	1.374474	0.476021	2.995649	0.056164	0.574915	0.000633	0.603657	0.555166
3.0007594	1.374289	0.476235	2.994898	0.056164	0.574914	0.000613	0.603824	0.555276
3.0001513	1.374112	0.47644	2.994177	0.056164	0.574913	0.000594	0.603984	0.555382
2.9995675	1.373942	0.476638	2.993485	0.056164	0.574913	0.000576	0.604138	0.555484
2.999007	1.373778	0.476827	2.992821	0.056164	0.574912	0.000559	0.604285	0.555581
2.9984689	1.373622	0.477009	2.992184	0.056164	0.574911	0.000543	0.604427	0.555675
2.9979522	1.373471	0.477183	2.991572	0.056164	0.57491	0.000527	0.604563	0.555765
Sce	nario 2 Mix	ing and Calcit	e Precipitatio	on				
Carbonate alkalinity	Ca++	CI-	HCO3-	K+	Mg++	NO3-	Na+	SO4
meq_acid/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l
3.6823246	1.584	0.2331	3.847656	0.05627	0.576	0.02278	0.4143	0.42995
3.6523582	1.573891	0.243031	3.811282	0.056266	0.575956	0.021875	0.422041	0.435069
3.619642	1.562154	0.252564	3.774331	0.056262	0.575913	0.021005	0.429473	0.439983
3.5881375	1.550839	0.261717	3.738809	0.056258	0.575872	0.020171	0.436607	0.444701
3.5578036	1.539932	0.270503	3.704665	0.056254	0.575833	0.01937	0.443456	0.44923
3.5286004	1.529421	0.278937	3.671846	0.05625	0.575796	0.018601	0.45003	0.453578
3.500489	1.519293	0.287035	3.640302	0.056247	0.575759	0.017863	0.456342	0.457751
3.4734315	1.509535	0.294808	3.609986	0.056243	0.575725	0.017154	0.462402	0.461758
3.4473914	1.500137	0.302271	3.58085	0.05624	0.575692	0.016474	0.468219	0.465605
3.4223328	1.491084	0.309435	3.552851	0.056237	0.57566	0.015821	0.473803	0.469298
3.3982213	1.482367	0.316312	3.525944	0.056234	0.575629	0.015193	0.479164	0.472843
3.3750231	1.473973	0.322915	3.500088	0.056231	0.575599	0.014592	0.484311	0.476246
3.3527057	1							
	1.465892	0.329253	3.475244	0.056228	0.575571	0.014014	0.489251	0.479513
3.3312377	1.465892 1.458113	0.329253 0.335338	3.475244 3.451372	0.056228 0.056225	0.575571 0.575544	0.014014 0.013459	0.489251 0.493995	0.479513 0.48265
3.3312377 3.3105882	1.465892 1.458113 1.450625	0.329253 0.335338 0.341179	3.475244 3.451372 3.428435	0.056228 0.056225 0.056223	0.575571 0.575544 0.575518	0.014014 0.013459 0.012926	0.489251 0.493995 0.498548	0.479513 0.48265 0.485661
3.3312377 3.3105882 3.2907279	1.465892 1.458113 1.450625 1.443419	0.329253 0.335338 0.341179 0.346787	3.475244 3.451372 3.428435 3.406398	0.056228 0.056225 0.056223 0.05622	0.575571 0.575544 0.575518 0.575493	0.014014 0.013459 0.012926 0.012415	0.489251 0.493995 0.498548 0.502919	0.479513 0.48265 0.485661 0.488551

3.2532604	1.429812	0.357338	3.364884	0.056216	0.575446	0.011453	0.511144	0.49399
3.2355988	1.423392	0.3623	3.345343	0.056214	0.575424	0.011001	0.515011	0.496548
3.2186169	1.417216	0.367063	3.32657	0.056212	0.575403	0.010567	0.518724	0.499003
3.2022897	1.411275	0.371635	3.308536	0.05621	0.575382	0.01015	0.522288	0.50136
3.1865929	1.405561	0.376024	3.291212	0.056208	0.575363	0.00975	0.52571	0.503622
3.1715031	1.400065	0.380238	3.274571	0.056206	0.575344	0.009365	0.528995	0.505794
3.1569977	1.394779	0.384284	3.258586	0.056204	0.575326	0.008996	0.532148	0.50788
3.1430547	1.389696	0.388167	3.243232	0.056202	0.575309	0.008642	0.535175	0.509882
3.129653	1.384808	0.391896	3.228484	0.056201	0.575292	0.008303	0.538082	0.511803
3.1167724	1.380109	0.395475	3.214318	0.056199	0.575276	0.007976	0.540871	0.513648
3.1043931	1.37559	0.398911	3.200713	0.056198	0.575261	0.007663	0.54355	0.515419
3.0924962	1.371246	0.402209	3.187645	0.056196	0.575246	0.007362	0.546121	0.51712
3.0810634	1.36707	0.405376	3.175094	0.056195	0.575232	0.007074	0.548589	0.518752
3.070077	1.363056	0.408416	3.16304	0.056194	0.575218	0.006796	0.550959	0.520319
3.0595201	1.359198	0.411334	3.151464	0.056192	0.575205	0.00653	0.553234	0.521823
3.0493762	1.355489	0.414135	3.140345	0.056191	0.575193	0.006275	0.555418	0.523267
3.0396296	1.351925	0.416825	3.129668	0.05619	0.575181	0.00603	0.557514	0.524653
3.030265	1.348499	0.419407	3.119413	0.056189	0.575169	0.005794	0.559527	0.525984
3.0212678	1.345207	0.421885	3.109566	0.056188	0.575158	0.005568	0.561459	0.527262
3.0126238	1.342043	0.424265	3.100109	0.056187	0.575148	0.005351	0.563314	0.528489
3.0043194	1.339003	0.426549	3.091027	0.056186	0.575137	0.005143	0.565094	0.529666
2.9963416	1.336081	0.428742	3.082305	0.056185	0.575128	0.004943	0.566804	0.530796
2.9886777	1.333274	0.430847	3.073931	0.056184	0.575118	0.004751	0.568445	0.531882
2.9813156	1.330577	0.432868	3.065888	0.056183	0.575109	0.004567	0.57002	0.532923
2.9742435	1.327986	0.434809	3.058165	0.056182	0.575101	0.00439	0.571532	0.533923
2.9674503	1.325496	0.436671	3.05075	0.056181	0.575092	0.00422	0.572984	0.534884
2.960925	1.323104	0.438459	3.043629	0.056181	0.575084	0.004057	0.574378	0.535805
2.9546573	1.320806	0.440176	3.036791	0.05618	0.575077	0.003901	0.575716	0.53669
2.9486371	1.318598	0.441824	3.030225	0.056179	0.575069	0.003751	0.577001	0.537539
2.9428548	1.316477	0.443406	3.02392	0.056178	0.575062	0.003606	0.578234	0.538355
2.9373011	1.31444	0.444924	3.017866	0.056178	0.575056	0.003468	0.579418	0.539138
2.931967	1.312483	0.446382	3.012053	0.056177	0.575049	0.003335	0.580554	0.539889
2.9268441	1.310603	0.447782	3.006472	0.056176	0.575043	0.003207	0.581645	0.540611
2.9219239	1.308798	0.449126	3.001112	0.056176	0.575037	0.003085	0.582693	0.541303
2.9171986	1.307063	0.450416	2.995966	0.056175	0.575031	0.002967	0.583698	0.541968
2.9126605	1.305397	0.451654	2.991026	0.056175	0.575026	0.002854	0.584663	0.542606
2.9083023	1.303797	0.452843	2.986281	0.056174	0.57502	0.002746	0.58559	0.543219
2.9041168	1.302261	0.453984	2.981726	0.056174	0.575015	0.002642	0.58648	0.543808
2.9000974	1.300785	0.455079	2.977353	0.056173	0.57501	0.002542	0.587334	0.544372
2.8962375	1.299367	0.456131	2.973154	0.056173	0.575006	0.002446	0.588153	0.544914
2.8925308	1.298006	0.457141	2.969122	0.056172	0.575001	0.002354	0.58894	0.545435
2.8889713	1.296698	0.45811	2.965251	0.056172	0.574997	0.002266	0.589696	0.545934
2.8855531	1.295442	0.459041	2.961534	0.056172	0.574993	0.002181	0.590421	0.546414

2.8822707	1.294236	0.459934	2.957965	0.056171	0.574989	0.002099	0.591118	0.546875
2.8791188	1.293078	0.460792	2.954539	0.056171	0.574985	0.002021	0.591786	0.547317
2.8760922	1.291966	0.461615	2.95125	0.05617	0.574981	0.001946	0.592428	0.547741
2.8731858	1.290898	0.462405	2.948091	0.05617	0.574978	0.001874	0.593044	0.548148
2.8703951	1.289872	0.463164	2.945059	0.05617	0.574974	0.001805	0.593635	0.548539
2.8677154	1.288887	0.463892	2.942148	0.056169	0.574971	0.001739	0.594203	0.548915
2.8651423	1.287941	0.464591	2.939352	0.056169	0.574968	0.001675	0.594748	0.549275
2.8626716	1.287033	0.465263	2.936669	0.056169	0.574965	0.001614	0.595272	0.549621
2.8602992	1.286161	0.465907	2.934092	0.056169	0.574962	0.001555	0.595774	0.549954
2.8580213	1.285323	0.466526	2.931619	0.056168	0.574959	0.001498	0.596256	0.550272
2.8558341	1.284519	0.46712	2.929244	0.056168	0.574957	0.001444	0.596719	0.550579
2.853734	1.283747	0.46769	2.926964	0.056168	0.574954	0.001392	0.597163	0.550872
2.8517176	1.283006	0.468237	2.924775	0.056168	0.574952	0.001342	0.59759	0.551155
2.8497815	1.282294	0.468763	2.922673	0.056167	0.574949	0.001295	0.598	0.551425
2.8479225	1.28161	0.469267	2.920655	0.056167	0.574947	0.001249	0.598393	0.551685
2.8461377	1.280953	0.469751	2.918718	0.056167	0.574945	0.001204	0.59877	0.551935
2.8444239	1.280323	0.470216	2.916858	0.056167	0.574943	0.001162	0.599133	0.552175
2.8427785	1.279718	0.470662	2.915073	0.056167	0.574941	0.001121	0.599481	0.552405
2.8411987	1.279137	0.471091	2.913358	0.056166	0.574939	0.001082	0.599815	0.552626
2.8396819	1.278579	0.471502	2.911713	0.056166	0.574937	0.001045	0.600135	0.552838
2.8382256	1.278043	0.471897	2.910133	0.056166	0.574935	0.001009	0.600443	0.553041
2.8368274	1.277529	0.472276	2.908616	0.056166	0.574934	0.000974	0.600738	0.553236
2.8354849	1.277035	0.47264	2.90716	0.056166	0.574932	0.000941	0.601022	0.553424
2.834196	1.276561	0.472989	2.905761	0.056166	0.57493	0.000909	0.601294	0.553604
2.8329585	1.276105	0.473325	2.904419	0.056165	0.574929	0.000879	0.601556	0.553777
2.8317704	1.275668	0.473647	2.903131	0.056165	0.574927	0.000849	0.601807	0.553943
2.8306297	1.275248	0.473956	2.901893	0.056165	0.574926	0.000821	0.602048	0.554102
2.8295346	1.274846	0.474252	2.900706	0.056165	0.574925	0.000794	0.602279	0.554255
2.8284831	1.274459	0.474537	2.899566	0.056165	0.574923	0.000768	0.602501	0.554402
2.8274737	1.274087	0.474811	2.898471	0.056165	0.574922	0.000743	0.602714	0.554543
2.8265045	1.273731	0.475073	2.89742	0.056165	0.574921	0.000719	0.602919	0.554678
2.825574	1.273388	0.475325	2.896411	0.056164	0.57492	0.000696	0.603115	0.554808
2.8246807	1.273059	0.475567	2.895443	0.056164	0.574919	0.000674	0.603304	0.554933
2.823823	1.272744	0.475799	2.894513	0.056164	0.574918	0.000653	0.603485	0.555053
2.8229996	1.272441	0.476022	2.89362	0.056164	0.574917	0.000633	0.603659	0.555167
2.8222091	1.27215	0.476236	2.892763	0.056164	0.574916	0.000613	0.603825	0.555278
2.8214501	1.27187	0.476442	2.89194	0.056164	0.574915	0.000594	0.603986	0.555384
2.8207215	1.271602	0.476639	2.89115	0.056164	0.574914	0.000576	0.604139	0.555485
2.8200219	1.271345	0.476828	2.890392	0.056164	0.574913	0.000559	0.604287	0.555583
2.8193503	1.271098	0.47701	2.889664	0.056164	0.574912	0.000543	0.604429	0.555677
2.8187056	1.27086	0.477185	2.888965	0.056164	0.574912	0.000527	0.604565	0.555767
Scenario	3 Mixing a	nd Cation Exc	change					

Carbonate alkalinity	Ca++	CI-	HCO3-	K+	Mg++	NO3-	Na+	SO4
meq_acid/l	mmol/l							
3.6823246	1.584	0.2331	3.847656	0.05627	0.576	0.02278	0.4143	0.42995
3.6553197	1.579961	0.243031	3.812825	0.056197	0.574551	0.021875	0.415865	0.435069
3.6293684	1.576066	0.252564	3.779388	0.056126	0.573152	0.021005	0.417423	0.439983
3.6044278	1.572307	0.261717	3.747288	0.056057	0.5718	0.020171	0.418974	0.444701
3.580457	1.56868	0.270503	3.716472	0.05599	0.570493	0.01937	0.420517	0.44923
3.5574167	1.56518	0.278937	3.686889	0.055926	0.569231	0.018601	0.422052	0.453577
3.5352695	1.561801	0.287035	3.658489	0.055863	0.568011	0.017863	0.423579	0.457751
3.5139791	1.558541	0.294808	3.631225	0.055803	0.566831	0.017154	0.425099	0.461758
3.493511	1.555392	0.302271	3.605051	0.055744	0.56569	0.016474	0.42661	0.465605
3.4738323	1.552353	0.309435	3.579925	0.055687	0.564586	0.015821	0.428113	0.469297
3.4549112	1.549417	0.316312	3.555803	0.055632	0.563518	0.015193	0.429607	0.472843
3.4367175	1.546582	0.322914	3.532646	0.055579	0.562485	0.014592	0.431093	0.476246
3.4192222	1.543844	0.329253	3.510416	0.055527	0.561484	0.014014	0.43257	0.479513
3.4023977	1.541198	0.335337	3.489075	0.055476	0.560516	0.013459	0.434038	0.482649
3.3862173	1.538641	0.341179	3.468587	0.055427	0.559579	0.012926	0.435497	0.48566
3.3706557	1.536169	0.346786	3.448919	0.05538	0.558671	0.012415	0.436948	0.488551
3.3556889	1.53378	0.35217	3.430038	0.055334	0.557792	0.011924	0.438389	0.491326
3.3412934	1.531471	0.357338	3.411911	0.055289	0.55694	0.011453	0.439822	0.49399
3.3274473	1.529237	0.362299	3.39451	0.055246	0.556115	0.011001	0.441246	0.496547
3.3141293	1.527076	0.367062	3.377805	0.055203	0.555316	0.010567	0.44266	0.499002
3.3013192	1.524985	0.371634	3.361768	0.055163	0.554541	0.01015	0.444066	0.501359
3.2889975	1.522962	0.376024	3.346373	0.055123	0.553791	0.009749	0.445462	0.503622
3.2771457	1.521003	0.380238	3.331593	0.055084	0.553063	0.009365	0.44685	0.505794
3.265746	1.519106	0.384283	3.317405	0.055047	0.552357	0.008996	0.448228	0.507879
3.2547812	1.51727	0.388167	3.303784	0.05501	0.551674	0.008642	0.449597	0.509881
3.2442352	1.51549	0.391895	3.290708	0.054975	0.551011	0.008303	0.450957	0.511802
3.2340921	1.513765	0.395474	3.278155	0.054941	0.550368	0.007976	0.452309	0.513647
3.2243368	1.512094	0.39891	3.266104	0.054908	0.549744	0.007663	0.453651	0.515418
3.214955	1.510473	0.402208	3.254535	0.054875	0.54914	0.007362	0.454984	0.517119
3.2059327	1.508901	0.405375	3.243429	0.054844	0.548554	0.007074	0.456309	0.518751
3.1972565	1.507375	0.408415	3.232767	0.054814	0.547985	0.006796	0.457624	0.520318
3.1889135	1.505895	0.411333	3.222532	0.054785	0.547433	0.00653	0.458931	0.521822
3.1808913	1.504458	0.414135	3.212706	0.054756	0.546899	0.006275	0.460228	0.523266
3.173178	1.503062	0.416824	3.203273	0.054728	0.54638	0.00603	0.461517	0.524652
3.1657622	1.501706	0.419406	3.194217	0.054702	0.545876	0.005794	0.462797	0.525983
3.1586326	1.500388	0.421885	3.185524	0.054676	0.545388	0.005568	0.464068	0.527261
3.1517786	1.499107	0.424264	3.177178	0.054651	0.544914	0.005351	0.46533	0.528487
3.14519	1.497862	0.426548	3.169166	0.054626	0.544454	0.005143	0.466584	0.529665
3.1388567	1.49665	0.428741	3.161475	0.054603	0.544008	0.004943	0.467828	0.530795
3.1327692	1.495472	0.430846	3.154091	0.05458	0.543575	0.004751	0.469064	0.53188

3.1269184	1.494324	0.432867	3.147003	0.054558	0.543155	0.004567	0.470292	0.532922
3.1212952	1.493207	0.434808	3.140198	0.054536	0.542747	0.00439	0.47151	0.533922
3.115891	1.492119	0.43667	3.133665	0.054515	0.542351	0.00422	0.47272	0.534882
3.1106977	1.49106	0.438458	3.127394	0.054495	0.541967	0.004057	0.473921	0.535804
3.1057072	1.490027	0.440175	3.121373	0.054476	0.541593	0.003901	0.475114	0.536689
3.1009119	1.48902	0.441823	3.115594	0.054457	0.541231	0.003751	0.476297	0.537538
3.0963042	1.488038	0.443405	3.110045	0.054439	0.540879	0.003606	0.477472	0.538354
3.091877	1.48708	0.444923	3.104719	0.054421	0.540537	0.003468	0.478639	0.539136
3.0876235	1.486146	0.446381	3.099605	0.054404	0.540205	0.003335	0.479797	0.539888
3.0835371	1.485233	0.447781	3.094696	0.054388	0.539883	0.003207	0.480946	0.540609
3.0796112	1.484343	0.449125	3.089983	0.054372	0.539569	0.003085	0.482087	0.541302
3.0758397	1.483473	0.450414	3.085459	0.054357	0.539265	0.002967	0.483219	0.541967
3.0722167	1.482624	0.451653	3.081116	0.054342	0.538969	0.002854	0.484342	0.542605
3.0687365	1.481793	0.452841	3.076947	0.054327	0.538681	0.002746	0.485457	0.543218
3.0653935	1.480982	0.453983	3.072944	0.054314	0.538402	0.002642	0.486563	0.543806
3.0621825	1.480188	0.455078	3.069102	0.0543	0.53813	0.002542	0.487661	0.544371
3.0590983	1.479412	0.45613	3.065413	0.054287	0.537866	0.002446	0.488751	0.544913
3.056136	1.478653	0.45714	3.061872	0.054275	0.537609	0.002354	0.489832	0.545433
3.0532908	1.47791	0.458109	3.058472	0.054262	0.537359	0.002266	0.490904	0.545933
3.0505583	1.477183	0.459039	3.055208	0.054251	0.537115	0.002181	0.491968	0.546413
3.047934	1.476471	0.459933	3.052075	0.054239	0.536879	0.002099	0.493024	0.546873
3.0454137	1.475773	0.46079	3.049067	0.054228	0.536648	0.002021	0.494071	0.547315
3.0429933	1.47509	0.461614	3.04618	0.054218	0.536424	0.001946	0.49511	0.54774
3.0406689	1.474421	0.462404	3.043408	0.054208	0.536206	0.001874	0.49614	0.548147
3.0384368	1.473765	0.463163	3.040747	0.054198	0.535994	0.001805	0.497162	0.548538
3.0362933	1.473122	0.463891	3.038192	0.054188	0.535787	0.001739	0.498176	0.548913
3.0342351	1.472491	0.46459	3.03574	0.054179	0.535586	0.001675	0.499182	0.549274
3.0322586	1.471872	0.465261	3.033385	0.05417	0.535389	0.001614	0.50018	0.54962
3.0303608	1.471265	0.465906	3.031125	0.054161	0.535198	0.001555	0.501169	0.549952
3.0285384	1.47067	0.466524	3.028955	0.054153	0.535012	0.001498	0.50215	0.550271
3.0267887	1.470085	0.467118	3.026872	0.054145	0.534831	0.001444	0.503123	0.550577
3.0251086	1.469511	0.467688	3.024873	0.054137	0.534654	0.001392	0.504088	0.550871
3.0234954	1.468947	0.468236	3.022953	0.05413	0.534482	0.001342	0.505045	0.551153
3.0219466	1.468394	0.468761	3.02111	0.054123	0.534314	0.001295	0.505994	0.551424
3.0204595	1.46785	0.469266	3.019341	0.054116	0.53415	0.001249	0.506935	0.551684
3.0190317	1.467315	0.46975	3.017643	0.054109	0.53399	0.001204	0.507867	0.551934
3.0176609	1.46679	0.470215	3.016012	0.054103	0.533834	0.001162	0.508793	0.552173
3.0163448	1.466273	0.470661	3.014447	0.054096	0.533682	0.001121	0.50971	0.552403
3.0150812	1.465765	0.471089	3.012944	0.05409	0.533534	0.001082	0.510619	0.552624
3.0138682	1.465266	0.471501	3.011502	0.054085	0.533389	0.001045	0.511521	0.552836
3.0127036	1.464774	0.471896	3.010117	0.054079	0.533248	0.001009	0.512415	0.55304
3.0115855	1.464291	0.472275	3.008787	0.054074	0.53311	0.000974	0.513301	0.553235
3.0105122	1.463815	0.472639	3.007511	0.054069	0.532975	0.000941	0.514179	0.553422

3.0094818	1.463347	0.472988	3.006286	0.054064	0.532844	0.000909	0.51505	0.553603
3.0084926	1.462886	0.473323	3.00511	0.054059	0.532716	0.000879	0.515913	0.553775
3.007543	1.462432	0.473645	3.003981	0.054054	0.53259	0.000849	0.516769	0.553941
3.0066314	1.461985	0.473954	3.002897	0.05405	0.532468	0.000821	0.517618	0.554101
3.0057564	1.461544	0.474251	3.001856	0.054045	0.532348	0.000794	0.518459	0.554254
3.0049163	1.46111	0.474536	3.000857	0.054041	0.532231	0.000768	0.519292	0.5544
3.00411	1.460683	0.474809	2.999898	0.054037	0.532117	0.000743	0.520118	0.554541
3.0033359	1.460262	0.475072	2.998977	0.054033	0.532006	0.000719	0.520937	0.554677
3.0025929	1.459846	0.475324	2.998093	0.05403	0.531896	0.000696	0.521749	0.554807
3.0018797	1.459437	0.475566	2.997245	0.054026	0.53179	0.000674	0.522553	0.554931
3.0011951	1.459033	0.475798	2.99643	0.054023	0.531686	0.000653	0.523351	0.555051
3.000538	1.458635	0.476021	2.995648	0.05402	0.531584	0.000633	0.524141	0.555166
2.9999073	1.458243	0.476235	2.994898	0.054016	0.531484	0.000613	0.524924	0.555276
2.9993018	1.457856	0.47644	2.994177	0.054013	0.531386	0.000594	0.5257	0.555382
2.9987207	1.457474	0.476638	2.993485	0.054011	0.531291	0.000576	0.52647	0.555484
2.998163	1.457097	0.476827	2.992821	0.054008	0.531198	0.000559	0.527232	0.555581
2.9976276	1.456725	0.477009	2.992183	0.054005	0.531107	0.000543	0.527988	0.555675
2.9971138	1.456357	0.477183	2.991571	0.054003	0.531017	0.000527	0.528736	0.555765
Scena	rio 4 Mixin	g, cation exch	ange, and cal	cite precipita	tion			
Carbonate alkalinity	Ca++	CI-	HCO3-	K+	Mg++	NO3-	Na+	SO4
meq_acid/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l
3.6823246	1.584	0.2331	3.847656	0.05627	0.576	0.02278	0.4143	0.42995
3.6514762	1.578526	0.243031	3.810842	0.056176	0.574072	0.021875	0.415747	0.435069
3.6178369	1.57175	0.252564	3.773428	0.056064	0.571715	0.021005	0.417068	0.439983
3.5854701	1.565197	0.261717	3.737475	0.055956	0.569438	0.020171	0.41839	0.444701
3.5543306	1.558862	0.270503	3.702926	0.055852	0.567238	0.01937	0.419711	0.44923
3.5243746	1.552736	0.278937	3.669729	0.05575	0.565113	0.018601	0.421033	0.453578
3.4955596	1.546813	0.287035	3.637832	0.055653	0.56306	0.017863	0.422355	0.457751
3.4678445	1.541087	0.294808	3.607185	0.055558	0.561078	0.017154	0.423676	0.461758
3.4411894	1.535552	0.302271	3.57774	0.055467	0.559163	0.016474	0.424997	0.465605
3.415556	1.530201	0.309435	3.54945	0.055379	0.557315	0.015821	0.426317	0.469298
3.3909068	1.525028	0.316312	3.522272	0.055294	0.55553	0.015193	0.427637	0.472843
3.3672058	1.520028	0.322915	3.496163	0.055212	0.553807	0.014592	0.428955	0.476246
3.3444181	1.515194	0.329253	3.471081	0.055132	0.552144	0.014014	0.430272	0.479513
3.3225101	1.510522	0.335338	3.446986	0.055056	0.550538	0.013459	0.431587	0.48265
3.3014491	1.506006	0.341179	3.423841	0.054982	0.548988	0.012926	0.432901	0.485661
3.2812037	1.50164	0.346787	3.401608	0.05491	0.547491	0.012415	0.434213	0.488551
3.2617435	1.497419	0.35217	3.380253	0.054841	0.546047	0.011924	0.435522	0.491326
3.2430391	1.493339	0.357338	3.35974	0.054775	0.544653	0.011453	0.43683	0.49399
3.2250622	1.489395	0.3623	3.340038	0.054711	0.543307	0.011001	0.438134	0.496548
3.2077854	1.485581	0.367063	3.321115	0.054649	0.542008	0.010567	0.439436	0.499003

3.1911824	1.481894	0.371635	3.30294	0.054589	0.540755	0.01015	0.440735	0.50136
3.1752276	1.478329	0.376025	3.285485	0.054532	0.539545	0.00975	0.442031	0.503622
3.1598966	1.474882	0.380238	3.26872	0.054476	0.538377	0.009365	0.443323	0.505794
3.1451657	1.471548	0.384284	3.25262	0.054423	0.53725	0.008996	0.444612	0.50788
3.131012	1.468323	0.388167	3.237158	0.054371	0.536162	0.008642	0.445897	0.509882
3.1174135	1.465205	0.391896	3.222308	0.054321	0.535112	0.008303	0.447178	0.511803
3.104349	1.462188	0.395475	3.208048	0.054273	0.534098	0.007976	0.448455	0.513648
3.0917981	1.45927	0.398911	3.194354	0.054227	0.533119	0.007663	0.449727	0.515419
3.079741	1.456446	0.402209	3.181204	0.054183	0.532175	0.007362	0.450995	0.51712
3.0681588	1.453714	0.405376	3.168576	0.05414	0.531263	0.007074	0.452258	0.518752
3.0570333	1.45107	0.408416	3.15645	0.054098	0.530383	0.006796	0.453517	0.520319
3.0463467	1.448511	0.411334	3.144806	0.054059	0.529533	0.00653	0.45477	0.521823
3.0360822	1.446034	0.414135	3.133625	0.05402	0.528713	0.006275	0.456018	0.523267
3.0262235	1.443636	0.416825	3.122889	0.053983	0.527921	0.00603	0.457261	0.524653
3.0167548	1.441314	0.419407	3.11258	0.053947	0.527156	0.005794	0.458499	0.525984
3.007661	1.439065	0.421885	3.102682	0.053913	0.526417	0.005568	0.459731	0.527262
2.9989276	1.436887	0.424265	3.093178	0.05388	0.525704	0.005351	0.460957	0.528489
2.9905405	1.434778	0.426549	3.084053	0.053848	0.525015	0.005143	0.462177	0.529666
2.9824862	1.432734	0.428742	3.075291	0.053817	0.52435	0.004943	0.463392	0.530796
2.9747519	1.430753	0.430847	3.066879	0.053788	0.523707	0.004751	0.4646	0.531882
2.9673249	1.428834	0.432868	3.058802	0.053759	0.523087	0.004567	0.465802	0.532923
2.9601934	1.426973	0.434809	3.051048	0.053732	0.522487	0.00439	0.466998	0.533923
2.9533457	1.42517	0.436671	3.043603	0.053706	0.521908	0.00422	0.468188	0.534884
2.9467707	1.42342	0.438459	3.036455	0.05368	0.521348	0.004057	0.469371	0.535805
2.9404579	1.421724	0.440176	3.029593	0.053656	0.520807	0.003901	0.470547	0.53669
2.9343968	1.420078	0.441824	3.023005	0.053632	0.520285	0.003751	0.471717	0.537539
2.9285776	1.418481	0.443406	3.01668	0.05361	0.51978	0.003606	0.47288	0.538355
2.9229907	1.416932	0.444924	3.010608	0.053588	0.519291	0.003468	0.474036	0.539138
2.9176272	1.415428	0.446382	3.004778	0.053567	0.51882	0.003335	0.475185	0.539889
2.9124781	1.413968	0.447782	2.999182	0.053547	0.518363	0.003207	0.476327	0.540611
2.9075351	1.41255	0.449126	2.99381	0.053527	0.517922	0.003085	0.477462	0.541303
2.90279	1.411172	0.450416	2.988653	0.053509	0.517496	0.002967	0.47859	0.541968
2.898235	1.409834	0.451654	2.983702	0.053491	0.517083	0.002854	0.479711	0.542606
2.8938626	1.408534	0.452843	2.97895	0.053473	0.516685	0.002746	0.480825	0.543219
2.8896656	1.407271	0.453984	2.974388	0.053457	0.516299	0.002642	0.481931	0.543808
2.8856371	1.406042	0.455079	2.970008	0.053441	0.515926	0.002542	0.483031	0.544372
2.8817704	1.404848	0.456131	2.965804	0.053425	0.515565	0.002446	0.484122	0.544914
2.8780591	1.403686	0.457141	2.961769	0.053411	0.515215	0.002354	0.485207	0.545435
2.874497	1.402556	0.45811	2.957895	0.053396	0.514877	0.002266	0.486284	0.545934
2.8710783	1.401456	0.459041	2.954177	0.053383	0.51455	0.002181	0.487354	0.546414
2.8677973	1.400386	0.459934	2.950608	0.05337	0.514234	0.002099	0.488416	0.546875
2.8646485	1.399344	0.460792	2.947183	0.053357	0.513927	0.002021	0.489471	0.547317
2.8616266	1.398329	0.461615	2.943894	0.053345	0.513631	0.001946	0.490518	0.547741

2.8587267	1.397341	0.462405	2.940738	0.053333	0.513343	0.001874	0.491558	0.548148
2.8559438	1.396378	0.463164	2.937709	0.053322	0.513065	0.001805	0.49259	0.548539
2.8532734	1.395439	0.463892	2.934801	0.053311	0.512796	0.001739	0.493615	0.548915
2.8507109	1.394525	0.464591	2.932011	0.053301	0.512535	0.001675	0.494632	0.549275
2.8482521	1.393633	0.465263	2.929332	0.053291	0.512282	0.001614	0.495642	0.549621
2.8458929	1.392763	0.465907	2.926761	0.053282	0.512037	0.001555	0.496644	0.549954
2.8436293	1.391915	0.466526	2.924294	0.053273	0.5118	0.001498	0.497638	0.550272
2.8414574	1.391087	0.46712	2.921926	0.053264	0.51157	0.001444	0.498625	0.550579
2.8393736	1.390279	0.46769	2.919653	0.053256	0.511347	0.001392	0.499605	0.550872
2.8373745	1.38949	0.468237	2.917472	0.053248	0.511131	0.001342	0.500577	0.551155
2.8354566	1.38872	0.468763	2.915379	0.05324	0.510922	0.001295	0.501541	0.551425
2.8336167	1.387967	0.469267	2.91337	0.053233	0.510719	0.001249	0.502498	0.551685
2.8318517	1.387232	0.469751	2.911442	0.053226	0.510522	0.001204	0.503448	0.551935
2.8301586	1.386514	0.470216	2.909592	0.053219	0.510331	0.001162	0.50439	0.552175
2.8285345	1.385812	0.470662	2.907817	0.053213	0.510146	0.001121	0.505324	0.552405
2.8269768	1.385125	0.471091	2.906113	0.053206	0.509967	0.001082	0.506251	0.552626
2.8254826	1.384453	0.471502	2.904478	0.053201	0.509792	0.001045	0.507171	0.552838
2.8240495	1.383796	0.471897	2.902909	0.053195	0.509623	0.001009	0.508083	0.553041
2.8226751	1.383154	0.472276	2.901403	0.05319	0.50946	0.000974	0.508988	0.553236
2.8213569	1.382524	0.47264	2.899959	0.053185	0.5093	0.000941	0.509886	0.553424
2.8200928	1.381908	0.472989	2.898573	0.05318	0.509146	0.000909	0.510776	0.553604
2.8188806	1.381305	0.473325	2.897242	0.053175	0.508996	0.000879	0.511659	0.553777
2.8177182	1.380714	0.473646	2.895966	0.053171	0.508851	0.000849	0.512535	0.553943
2.8166036	1.380135	0.473956	2.894742	0.053166	0.508709	0.000821	0.513403	0.554102
2.8155349	1.379568	0.474252	2.893567	0.053162	0.508572	0.000794	0.514264	0.554255
2.8145103	1.379011	0.474537	2.89244	0.053159	0.508439	0.000768	0.515119	0.554402
2.813528	1.378466	0.474811	2.891358	0.053155	0.50831	0.000743	0.515965	0.554543
2.8125862	1.377931	0.475073	2.89032	0.053152	0.508184	0.000719	0.516805	0.554678
2.8116834	1.377406	0.475325	2.889325	0.053149	0.508062	0.000696	0.517638	0.554808
2.810818	1.376892	0.475567	2.88837	0.053146	0.507944	0.000674	0.518464	0.554933
2.8099886	1.376386	0.475799	2.887454	0.053143	0.507829	0.000653	0.519283	0.555053
2.8091935	1.37589	0.476022	2.886575	0.05314	0.507717	0.000633	0.520094	0.555167
2.8084316	1.375403	0.476236	2.885732	0.053137	0.507608	0.000613	0.520899	0.555278
2.8077014	1.374925	0.476442	2.884923	0.053135	0.507502	0.000594	0.521697	0.555384
2.8070017	1.374455	0.476639	2.884148	0.053133	0.5074	0.000576	0.522488	0.555485
2.8063312	1.373993	0.476828	2.883403	0.053131	0.5073	0.000559	0.523273	0.555583
2.8056888	1.373539	0.47701	2.88269	0.053129	0.507203	0.000543	0.52405	0.555677
2.8050733	1.373093	0.477185	2.882005	0.053127	0.507109	0.000527	0.524821	0.555767
Scer	ario 5 Mixi	ng and Catior	• Exchange *	10				
Carbonate	6			И.		NO2	N1 - 7	604
alkalinity	Ca++	CI-	нсоз-	К+	ı∨lg++	NU3-	Na+	504

meq_acid/l	mmol/l							
3.6823246	1.584	0.2331	3.847656	0.05627	0.576	0.02278	0.4143	0.42995
3.6553143	1.580657	0.243031	3.812825	0.056204	0.574747	0.021875	0.414075	0.435069
3.6293575	1.577449	0.252564	3.779388	0.056141	0.57354	0.021005	0.413866	0.439983
3.6044114	1.574368	0.261717	3.747288	0.05608	0.572378	0.020171	0.413672	0.444701
3.580435	1.571412	0.270503	3.716472	0.05602	0.571258	0.01937	0.413494	0.44923
3.557389	1.568573	0.278937	3.686889	0.055963	0.57018	0.018601	0.413329	0.453577
3.5352359	1.565849	0.287035	3.658489	0.055908	0.569141	0.017863	0.413178	0.457751
3.5139396	1.563234	0.294808	3.631225	0.055854	0.568141	0.017154	0.41304	0.461758
3.4934655	1.560725	0.302271	3.605051	0.055802	0.567176	0.016474	0.412915	0.465605
3.4737806	1.558316	0.309435	3.579925	0.055752	0.566246	0.015821	0.412801	0.469297
3.4548532	1.556003	0.316312	3.555803	0.055703	0.56535	0.015193	0.4127	0.472843
3.4366531	1.553784	0.322914	3.532646	0.055656	0.564486	0.014592	0.412609	0.476246
3.4191513	1.551654	0.329253	3.510416	0.05561	0.563652	0.014014	0.412529	0.479513
3.4023201	1.54961	0.335337	3.489075	0.055566	0.562848	0.013459	0.412459	0.482649
3.3861329	1.547648	0.341179	3.468587	0.055523	0.562073	0.012926	0.412399	0.48566
3.3705645	1.545764	0.346786	3.448919	0.055482	0.561325	0.012415	0.412349	0.488551
3.3555906	1.543956	0.35217	3.430038	0.055442	0.560603	0.011924	0.412307	0.491326
3.3411881	1.542221	0.357338	3.411912	0.055403	0.559907	0.011453	0.412275	0.49399
3.3273348	1.540555	0.362299	3.394511	0.055365	0.559235	0.011001	0.412251	0.496547
3.3140096	1.538956	0.367062	3.377805	0.055328	0.558586	0.010567	0.412236	0.499002
3.3011921	1.53742	0.371634	3.361769	0.055293	0.55796	0.01015	0.412228	0.501359
3.288863	1.535946	0.376024	3.346373	0.055258	0.557355	0.009749	0.412229	0.503622
3.2770037	1.53453	0.380238	3.331594	0.055225	0.556772	0.009365	0.412237	0.505794
3.2655964	1.53317	0.384283	3.317405	0.055192	0.556209	0.008996	0.412252	0.507879
3.254624	1.531864	0.388167	3.303784	0.055161	0.555666	0.008642	0.412274	0.509881
3.2440703	1.530609	0.391895	3.290708	0.055131	0.555141	0.008303	0.412304	0.511802
3.2339195	1.529403	0.395474	3.278155	0.055101	0.554634	0.007976	0.412339	0.513647
3.2241566	1.528245	0.39891	3.266104	0.055073	0.554145	0.007663	0.412382	0.515418
3.2147671	1.527132	0.402208	3.254535	0.055045	0.553673	0.007362	0.412431	0.517119
3.205737	1.526061	0.405375	3.243429	0.055018	0.553217	0.007074	0.412486	0.518751
3.1970531	1.525032	0.408415	3.232767	0.054992	0.552778	0.006796	0.412547	0.520318
3.1887023	1.524043	0.411333	3.222532	0.054967	0.552353	0.00653	0.412614	0.521822
3.1806724	1.523091	0.414135	3.212706	0.054943	0.551943	0.006275	0.412686	0.523266
3.1729515	1.522175	0.416824	3.203273	0.054919	0.551547	0.00603	0.412765	0.524652
3.1655279	1.521294	0.419406	3.194217	0.054897	0.551165	0.005794	0.412848	0.525983
3.1583906	1.520446	0.421885	3.185524	0.054875	0.550796	0.005568	0.412937	0.527261
3.151529	1.51963	0.424264	3.177178	0.054853	0.55044	0.005351	0.41303	0.528487
3.1449327	1.518843	0.426548	3.169166	0.054833	0.550097	0.005143	0.413129	0.529665
3.1385918	1.518086	0.428741	3.161475	0.054813	0.549765	0.004943	0.413233	0.530795
3.1324968	1.517357	0.430847	3.154091	0.054793	0.549445	0.004751	0.413341	0.53188
3.1266384	1.516654	0.432868	3.147003	0.054775	0.549136	0.004567	0.413453	0.532922
3.1210077	1.515977	0.434808	3.140198	0.054757	0.548837	0.00439	0.41357	0.533922

3.1155961	1.515324	0.43667	3.133665	0.054739	0.548549	0.00422	0.413691	0.534882
3.1103954	1.514694	0.438458	3.127394	0.054723	0.548271	0.004057	0.413816	0.535804
3.1053976	1.514087	0.440175	3.121374	0.054706	0.548003	0.003901	0.413945	0.536689
3.1005949	1.5135	0.441823	3.115594	0.054691	0.547743	0.003751	0.414078	0.537538
3.09598	1.512935	0.443405	3.110045	0.054675	0.547493	0.003606	0.414215	0.538354
3.0915457	1.512389	0.444923	3.104719	0.054661	0.547251	0.003468	0.414355	0.539136
3.0872851	1.511861	0.446381	3.099605	0.054646	0.547018	0.003335	0.414499	0.539888
3.0831915	1.511352	0.447781	3.094696	0.054633	0.546792	0.003207	0.414646	0.540609
3.0792587	1.51086	0.449125	3.089984	0.05462	0.546574	0.003085	0.414796	0.541302
3.0754803	1.510384	0.450414	3.08546	0.054607	0.546364	0.002967	0.414949	0.541967
3.0718504	1.509924	0.451653	3.081117	0.054594	0.546161	0.002854	0.415105	0.542605
3.0683634	1.50948	0.452841	3.076947	0.054583	0.545964	0.002746	0.415264	0.543218
3.0650137	1.509049	0.453983	3.072944	0.054571	0.545774	0.002642	0.415426	0.543806
3.061796	1.508633	0.455078	3.069102	0.05456	0.545591	0.002542	0.41559	0.544371
3.0587052	1.50823	0.45613	3.065413	0.054549	0.545414	0.002446	0.415757	0.544913
3.0557364	1.50784	0.45714	3.061872	0.054539	0.545242	0.002354	0.415926	0.545434
3.0528848	1.507462	0.458109	3.058472	0.054529	0.545077	0.002266	0.416098	0.545933
3.0501458	1.507096	0.459039	3.055209	0.054519	0.544916	0.002181	0.416272	0.546413
3.0475152	1.506741	0.459933	3.052075	0.05451	0.544761	0.002099	0.416448	0.546873
3.0449886	1.506397	0.46079	3.049068	0.054501	0.544611	0.002021	0.416627	0.547315
3.0425621	1.506063	0.461614	3.04618	0.054492	0.544466	0.001946	0.416807	0.54774
3.0402316	1.505739	0.462404	3.043408	0.054483	0.544326	0.001874	0.416989	0.548147
3.0379935	1.505425	0.463163	3.040747	0.054475	0.54419	0.001805	0.417173	0.548538
3.0358441	1.50512	0.463891	3.038193	0.054467	0.544058	0.001739	0.417359	0.548914
3.0337799	1.504824	0.46459	3.03574	0.05446	0.543931	0.001675	0.417546	0.549274
3.0317976	1.504536	0.465261	3.033386	0.054452	0.543807	0.001614	0.417735	0.54962
3.029894	1.504256	0.465906	3.031126	0.054445	0.543688	0.001555	0.417926	0.549952
3.028066	1.503984	0.466525	3.028956	0.054438	0.543572	0.001498	0.418118	0.550271
3.0263106	1.503719	0.467118	3.026873	0.054432	0.54346	0.001444	0.418311	0.550577
3.024625	1.503461	0.467689	3.024873	0.054425	0.543351	0.001392	0.418506	0.550871
3.0230064	1.503211	0.468236	3.022953	0.054419	0.543245	0.001342	0.418702	0.551153
3.0214522	1.502966	0.468761	3.021111	0.054413	0.543143	0.001295	0.4189	0.551424
3.0199597	1.502729	0.469266	3.019341	0.054407	0.543044	0.001249	0.419098	0.551684
3.0185267	1.502497	0.46975	3.017643	0.054401	0.542947	0.001204	0.419298	0.551934
3.0171507	1.502271	0.470215	3.016013	0.054396	0.542853	0.001162	0.419499	0.552173
3.0158295	1.502051	0.470661	3.014447	0.054391	0.542763	0.001121	0.419701	0.552403
3.0145609	1.501836	0.47109	3.012945	0.054386	0.542674	0.001082	0.419903	0.552624
3.0133429	1.501626	0.471501	3.011502	0.054381	0.542588	0.001045	0.420107	0.552836
3.0121733	1.501421	0.471896	3.010117	0.054376	0.542505	0.001009	0.420311	0.55304
3.0110504	1.501221	0.472275	3.008788	0.054371	0.542424	0.000974	0.420517	0.553235
3.0099723	1.501025	0.472639	3.007512	0.054367	0.542345	0.000941	0.420723	0.553423
3.0089372	1.500834	0.472988	3.006286	0.054362	0.542268	0.000909	0.42093	0.553603
3.0079433	1.500647	0.473323	3.00511	0.054358	0.542194	0.000879	0.421137	0.553775

3.0069891	1.500463	0.473645	3.003981	0.054354	0.542121	0.000849	0.421345	0.553941
3.006073	1.500284	0.473954	3.002897	0.05435	0.54205	0.000821	0.421554	0.554101
3.0051935	1.500109	0.474251	3.001856	0.054346	0.541981	0.000794	0.421764	0.554254
3.004349	1.499937	0.474536	3.000857	0.054342	0.541914	0.000768	0.421974	0.5544
3.0035383	1.499769	0.474809	2.999898	0.054338	0.541848	0.000743	0.422184	0.554541
3.00276	1.499603	0.475072	2.998978	0.054335	0.541784	0.000719	0.422395	0.554677
3.0020127	1.499442	0.475324	2.998094	0.054331	0.541722	0.000696	0.422607	0.554807
3.0012954	1.499283	0.475566	2.997245	0.054328	0.541661	0.000674	0.422819	0.554931
3.0006066	1.499127	0.475798	2.996431	0.054325	0.541601	0.000653	0.423031	0.555051
2.9999454	1.498974	0.476021	2.995649	0.054322	0.541543	0.000633	0.423244	0.555166
2.9993107	1.498823	0.476235	2.994898	0.054318	0.541487	0.000613	0.423457	0.555276
2.9987013	1.498675	0.47644	2.994177	0.054315	0.541431	0.000594	0.42367	0.555382
2.9981163	1.49853	0.476638	2.993485	0.054312	0.541377	0.000576	0.423884	0.555484
2.9975547	1.498387	0.476827	2.992821	0.05431	0.541324	0.000559	0.424098	0.555582
2.9970156	1.498246	0.477009	2.992184	0.054307	0.541272	0.000543	0.424312	0.555675
2.996498	1.498108	0.477183	2.991572	0.054304	0.541221	0.000527	0.424527	0.555765
So	cenario 6 M	ixing, cation o	exchange, and	d calcite prec	iptiation * 10			
Carbonate alkalinity	Ca++	CI-	HCO3-	K+	Mg++	NO3-	Na+	SO4
meq_acid/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l	mmol/l
3.6823246	1.584	0.2331	3.847656	0.05627	0.576	0.02278	0.4143	0.42995
3.6513535	1.579199	0.243031	3.810781	0.056183	0.57425	0.021875	0.413918	0.435069
3.6175928	1.573109	0.252564	3.773307	0.056076	0.572064	0.021005	0.4134	0.439983
3.5851093	1.567234	0.261717	3.737297	0.055974	0.569955	0.020171	0.412909	0.444701
3.5538577	1.561568	0.270503	3.702693	0.055875	0.567922	0.01937	0.412442	0.44923
3.5237937	1.556103	0.278937	3.669443	0.055779	0.565961	0.018601	0.412	0.453578
3.4948747	1.550835	0.287035	3.637494	0.055686	0.56407	0.017863	0.411581	0.457751
3.4670593	1.545755	0.294808	3.606798	0.055597	0.562248	0.017154	0.411186	0.461758
3.4403074	1.540859	0.302271	3.577305	0.055511	0.560491	0.016474	0.410813	0.465605
3.4145804	1.536139	0.309435	3.548969	0.055428	0.558798	0.015821	0.410462	0.469298
3.3898407	1.531591	0.316312	3.521746	0.055347	0.557166	0.015193	0.410132	0.472843
3.3660521	1.527208	0.322915	3.495593	0.05527	0.555594	0.014592	0.409822	0.476246
3.3431796	1.522985	0.329253	3.470469	0.055195	0.55408	0.014014	0.409533	0.479513
3.3211892	1.518916	0.335338	3.446334	0.055123	0.552621	0.013459	0.409262	0.48265
3.3000482	1.514996	0.341179	3.423149	0.055054	0.551215	0.012926	0.40901	0.485661
3.2797251	1.51122	0.346787	3.400878	0.054987	0.549862	0.012415	0.408776	0.488551
3.2601892	1.507582	0.35217	3.379485	0.054922	0.548558	0.011924	0.408559	0.491326
3.2414112	1.504079	0.357338	3.358936	0.05486	0.547302	0.011453	0.408359	0.49399
3.2233625	1.500704	0.3623	3.339199	0.0548	0.546093	0.011001	0.408175	0.496548
3.2060157	1.497454	0.367063	3.320241	0.054742	0.544929	0.010567	0.408007	0.499003
3.1893443	1.494325	0.371635	3.302032	0.054686	0.543808	0.01015	0.407855	0.50136
3.1733228	1.491311	0.376025	3.284544	0.054633	0.542729	0.00975	0.407716	0.503622

3.1579265	1.488408	0.380238	3.267747	0.054581	0.54169	0.009365	0.407592	0.505794
3.1431316	1.485613	0.384284	3.251615	0.054531	0.54069	0.008996	0.407482	0.50788
3.1289153	1.482922	0.388167	3.236121	0.054483	0.539727	0.008642	0.407385	0.509882
3.1152555	1.480331	0.391896	3.221242	0.054437	0.5388	0.008303	0.4073	0.511803
3.1021308	1.477836	0.395475	3.206952	0.054392	0.537907	0.007976	0.407228	0.513648
3.0895209	1.475433	0.398911	3.193228	0.054349	0.537048	0.007663	0.407168	0.515419
3.0774058	1.47312	0.402209	3.180049	0.054308	0.536222	0.007362	0.407119	0.51712
3.0657667	1.470893	0.405376	3.167393	0.054268	0.535426	0.007074	0.407081	0.518752
3.0545851	1.468748	0.408416	3.155239	0.05423	0.53466	0.006796	0.407054	0.520319
3.0438435	1.466683	0.411334	3.143568	0.054193	0.533923	0.00653	0.407037	0.521823
3.0335248	1.464694	0.414135	3.13236	0.054158	0.533213	0.006275	0.40703	0.523267
3.0236127	1.46278	0.416825	3.121597	0.054123	0.53253	0.00603	0.407032	0.524653
3.0140914	1.460936	0.419407	3.111263	0.05409	0.531872	0.005794	0.407043	0.525984
3.0049457	1.459161	0.421886	3.101339	0.054059	0.531239	0.005568	0.407063	0.527262
2.9961611	1.457451	0.424265	3.091809	0.054028	0.53063	0.005351	0.407092	0.528489
2.9877236	1.455804	0.426549	3.082659	0.053999	0.530044	0.005143	0.407129	0.529666
2.9796196	1.454219	0.428742	3.073872	0.05397	0.529479	0.004943	0.407173	0.530796
2.9718361	1.452691	0.430848	3.065436	0.053943	0.528936	0.004751	0.407225	0.531882
2.9643606	1.451221	0.432869	3.057335	0.053917	0.528413	0.004567	0.407285	0.532923
2.9571811	1.449804	0.434809	3.049557	0.053892	0.52791	0.00439	0.407351	0.533923
2.9502861	1.448439	0.436671	3.042088	0.053867	0.527425	0.00422	0.407424	0.534884
2.9436643	1.447124	0.438459	3.034917	0.053844	0.526958	0.004057	0.407504	0.535805
2.9373051	1.445858	0.440176	3.028032	0.053822	0.526509	0.003901	0.407589	0.53669
2.9311982	1.444637	0.441824	3.021421	0.0538	0.526076	0.003751	0.407681	0.53754
2.9253337	1.443462	0.443406	3.015074	0.053779	0.525659	0.003606	0.407778	0.538355
2.9197021	1.442328	0.444924	3.00898	0.053759	0.525258	0.003468	0.407881	0.539138
2.9142942	1.441236	0.446382	3.003128	0.05374	0.524872	0.003335	0.407989	0.539889
2.9091013	1.440184	0.447782	2.99751	0.053721	0.5245	0.003207	0.408103	0.540611
2.9041148	1.43917	0.449126	2.992117	0.053703	0.524141	0.003085	0.408221	0.541303
2.8993267	1.438192	0.450416	2.986938	0.053686	0.523796	0.002967	0.408343	0.541968
2.8947291	1.437249	0.451654	2.981966	0.053669	0.523464	0.002854	0.408471	0.542607
2.8903145	1.43634	0.452843	2.977193	0.053653	0.523143	0.002746	0.408602	0.543219
2.8860757	1.435463	0.453984	2.97261	0.053638	0.522835	0.002642	0.408738	0.543808
2.8820058	1.434618	0.455079	2.96821	0.053623	0.522537	0.002542	0.408878	0.544372
2.8780981	1.433802	0.456131	2.963985	0.053608	0.522251	0.002446	0.409021	0.544914
2.8743461	1.433015	0.457141	2.95993	0.053595	0.521974	0.002354	0.409168	0.545435
2.8707438	1.432256	0.45811	2.956036	0.053581	0.521708	0.002266	0.409319	0.545935
2.8672852	1.431523	0.459041	2.952298	0.053569	0.521452	0.002181	0.409473	0.546414
2.8639646	1.430816	0.459934	2.94871	0.053556	0.521204	0.002099	0.40963	0.546875
2.8607766	1.430134	0.460792	2.945264	0.053544	0.520966	0.002021	0.40979	0.547317
2.8577158	1.429475	0.461615	2.941957	0.053533	0.520735	0.001946	0.409953	0.547741
2.8547774	1.428839	0.462405	2.938782	0.053522	0.520514	0.001874	0.410119	0.548148
2.8519563	1.428224	0.463164	2.935733	0.053511	0.5203	0.001805	0.410288	0.54854

2.8492481	1.42763	0.463892	2.932807	0.053501	0.520093	0.001739	0.410459	0.548915
2.8466481	1.427057	0.464592	2.929997	0.053491	0.519894	0.001675	0.410633	0.549275
2.8441521	1.426503	0.465263	2.9273	0.053482	0.519702	0.001614	0.410809	0.549621
2.841756	1.425967	0.465907	2.924711	0.053472	0.519516	0.001555	0.410988	0.549954
2.8394558	1.425449	0.466526	2.922226	0.053464	0.519337	0.001498	0.411168	0.550273
2.8372476	1.424948	0.46712	2.91984	0.053455	0.519164	0.001444	0.411351	0.550579
2.8351279	1.424464	0.46769	2.917549	0.053447	0.518997	0.001392	0.411535	0.550873
2.8330932	1.423996	0.468237	2.91535	0.053439	0.518835	0.001342	0.411722	0.551155
2.8311399	1.423542	0.468763	2.913239	0.053431	0.51868	0.001295	0.41191	0.551425
2.829265	1.423104	0.469267	2.911213	0.053424	0.518529	0.001249	0.4121	0.551686
2.8274653	1.422679	0.469751	2.909268	0.053417	0.518383	0.001204	0.412292	0.551935
2.8257377	1.422267	0.470216	2.907401	0.05341	0.518243	0.001162	0.412486	0.552175
2.8240795	1.421869	0.470662	2.905608	0.053403	0.518107	0.001121	0.41268	0.552405
2.8224879	1.421483	0.471091	2.903887	0.053397	0.517975	0.001082	0.412877	0.552626
2.8209602	1.421109	0.471502	2.902236	0.05339	0.517848	0.001045	0.413074	0.552838
2.8194938	1.420747	0.471897	2.90065	0.053384	0.517725	0.001009	0.413273	0.553041
2.8180864	1.420395	0.472276	2.899128	0.053379	0.517606	0.000974	0.413473	0.553237
2.8167355	1.420054	0.47264	2.897667	0.053373	0.517491	0.000941	0.413675	0.553424
2.815439	1.419723	0.472989	2.896265	0.053368	0.517379	0.000909	0.413877	0.553604
2.8141947	1.419402	0.473325	2.894919	0.053362	0.517271	0.000879	0.414081	0.553777
2.8130004	1.41909	0.473647	2.893627	0.053357	0.517166	0.000849	0.414285	0.553943
2.8118542	1.418788	0.473956	2.892387	0.053352	0.517065	0.000821	0.414491	0.554102
2.8107542	1.418494	0.474252	2.891196	0.053348	0.516966	0.000794	0.414697	0.554255
2.8096986	1.418208	0.474537	2.890053	0.053343	0.516871	0.000768	0.414904	0.554402
2.8086854	1.41793	0.474811	2.888957	0.053339	0.516779	0.000743	0.415112	0.554543
2.8077132	1.41766	0.475073	2.887904	0.053334	0.516689	0.000719	0.415321	0.554678
2.8067802	1.417397	0.475325	2.886893	0.05333	0.516602	0.000696	0.415531	0.554808
2.8058848	1.417142	0.475567	2.885923	0.053326	0.516517	0.000674	0.415741	0.554933
2.8050257	1.416893	0.475799	2.884992	0.053322	0.516435	0.000653	0.415952	0.555053
2.8042012	1.416651	0.476022	2.884099	0.053318	0.516356	0.000633	0.416164	0.555168
2.8034101	1.416415	0.476236	2.883241	0.053315	0.516278	0.000613	0.416376	0.555278
2.802651	1.416185	0.476442	2.882418	0.053311	0.516203	0.000594	0.416589	0.555384
2.8019226	1.415961	0.476639	2.881628	0.053308	0.51613	0.000576	0.416802	0.555485
2.8012238	1.415742	0.476828	2.88087	0.053304	0.516059	0.000559	0.417016	0.555583
2.8005532	1.415529	0.47701	2.880142	0.053301	0.51599	0.000543	0.41723	0.555677
2.7999099	1.415321	0.477185	2.879444	0.053298	0.515922	0.000527	0.417445	0.555767

Appendix H – Modeled changes in aquifer sediments for Scenario Six (mixing, cation

Mass reacted, H2O	Ca++	K+	Mg++	Na+	Mass reacted, H2O	Calcite_ESRPA	Porosity	0.41
kg	meq/l	meq/l	meq/l	meq/l	kg	mg	CEC (meq/g rock)	0.21
Mass reacted, H2O	Ca++	K+	Mg++	Na+				
0	0	0	0	0	0			
0.04	-0.0099	8.57E-05	0.003666	0.006996	0.04	0.1739		
0.08	-0.02101	0.000196	0.008483	0.014108	0.08	0.58267		
0.12	-0.03213	0.000308	0.01332	0.021182	0.12	0.995034		
0.16	-0.04326	0.00042	0.018175	0.028217	0.16	1.410715		
0.2	-0.05441	0.000532	0.023048	0.035215	0.2	1.829453		
0.24	-0.06556	0.000646	0.027938	0.042174	0.24	2.251009		
0.28	-0.07671	0.00076	0.032843	0.049094	0.28	2.675163		
0.32	-0.08787	0.000874	0.037763	0.055977	0.32	3.101708		
0.36	-0.09904	0.000989	0.042697	0.062821	0.36	3.530458		
0.4	-0.1102	0.001104	0.047645	0.069626	0.4	3.961235		
0.44	-0.12136	0.00122	0.052604	0.076394	0.44	4.393879		
0.48	-0.13252	0.001336	0.057576	0.083122	0.48	4.82824		
0.52	-0.14367	0.001453	0.062559	0.089813	0.52	5.26418		
0.56	-0.15482	0.00157	0.067554	0.096465	0.56	5.701571		
0.6	-0.16596	0.001687	0.072558	0.103079	0.6	6.140295		
0.64	-0.17709	0.001805	0.077573	0.109655	0.64	6.580244		
0.68	-0.18822	0.001923	0.082597	0.116193	0.68	7.021317		
0.72	-0.19933	0.002041	0.08763	0.122693	0.72	7.463421		
0.76	-0.21043	0.00216	0.092672	0.129155	0.76	7.90647		
0.8	-0.22152	0.002278	0.097723	0.135579	0.8	8.350386		
0.84	-0.2326	0.002397	0.102783	0.141966	0.84	8.795094		
0.88	-0.24366	0.002517	0.10785	0.148315	0.88	9.240529		
0.92	-0.25471	0.002636	0.112925	0.154626	0.92	9.686626		
0.96	-0.26574	0.002756	0.118008	0.1609	0.96	10.13333		
1	-0.27676	0.002876	0.123098	0.167137	1	10.58059		
1.04	-0.28776	0.002996	0.128195	0.173338	1.04	11.02835		
1.08	-0.29874	0.003116	0.133299	0.179501	1.08	11.47657		
1.12	-0.30971	0.003237	0.138411	0.185627	1.12	11.92521		
1.16	-0.32066	0.003358	0.143529	0.191717	1.16	12.37423		
1.2	-0.33159	0.003479	0.148653	0.197771	1.2	12.82359		

exchange, and calcite precipitation) at the alluvial site – The Jones Pit

1.24	-0.3425	0.0036	0.153784	0.203788	1.24	13.27326	
1.279999	-0.3534	0.003721	0.158922	0.209769	1.279999	13.72322	
1.319999	-0.36427	0.003842	0.164066	0.215714	1.319999	14.17343	
1.359999	-0.37512	0.003963	0.169216	0.221623	1.359999	14.62387	
1.399999	-0.38596	0.004085	0.174372	0.227497	1.399999	15.07451	
1.439999	-0.39677	0.004207	0.179534	0.233335	1.439999	15.52533	
1.479999	-0.40756	0.004329	0.184702	0.239138	1.479999	15.97632	
1.519999	-0.41833	0.004451	0.189875	0.244906	1.519999	16.42746	
1.559999	-0.42909	0.004573	0.195055	0.25064	1.559999	16.87872	
1.599999	-0.43981	0.004695	0.20024	0.256338	1.599999	17.33009	
1.639999	-0.45052	0.004817	0.205431	0.262002	1.639999	17.78156	
1.679999	-0.46121	0.00494	0.210628	0.267631	1.679999	18.23312	
1.719999	-0.47187	0.005062	0.21583	0.273226	1.719999	18.68474	
1.759999	-0.48252	0.005185	0.221037	0.278788	1.759999	19.13643	
1.799999	-0.49314	0.005308	0.22625	0.284315	1.799999	19.58817	
1.839999	-0.50373	0.00543	0.231469	0.289809	1.839999	20.03995	
1.879999	-0.51431	0.005553	0.236692	0.295269	1.879999	20.49177	
1.919999	-0.52486	0.005676	0.241921	0.300696	1.919999	20.94361	
1.959999	-0.53539	0.0058	0.247156	0.30609	1.959999	21.39546	
1.999999	-0.5459	0.005923	0.252395	0.311452	1.999999	21.84733	
2.039999	-0.55639	0.006046	0.25764	0.31678	2.039999	22.2992	
2.079999	-0.56685	0.006169	0.26289	0.322076	2.079999	22.75106	
2.119999	-0.57729	0.006293	0.268145	0.327339	2.119999	23.20293	
2.159999	-0.58771	0.006416	0.273405	0.332571	2.159999	23.65478	
2.199999	-0.59811	0.00654	0.27867	0.33777	2.199999	24.10661	
2.239999	-0.60848	0.006663	0.28394	0.342938	2.239999	24.55842	
2.279999	-0.61883	0.006787	0.289215	0.348074	2.279999	25.01021	
2.319999	-0.62916	0.006911	0.294495	0.353178	2.319999	25.46197	
2.359999	-0.63946	0.007035	0.299779	0.358252	2.359999	25.91371	
2.399999	-0.64974	0.007159	0.305069	0.363294	2.399999	26.3654	
2.439999	-0.66	0.007283	0.310364	0.368305	2.439999	26.81707	
2.479999	-0.67024	0.007407	0.315663	0.373286	2.479999	27.26869	
2.519999	-0.68045	0.007531	0.320967	0.378236	2.519999	27.72028	
2.559999	-0.69064	0.007655	0.326276	0.383156	2.559999	28.17182	
2.599999	-0.70081	0.007779	0.331589	0.388046	2.599999	28.62332	
2.639999	-0.71096	0.007903	0.336907	0.392906	2.639999	29.07477	
2.679999	-0.72108	0.008028	0.34223	0.397736	2.679999	29.52617	
2.719999	-0.73118	0.008152	0.347558	0.402536	2.719999	29.97753	
2.759999	-0.74126	0.008277	0.35289	0.407307	2.759999	30.42883	
2.799999	-0.75131	0.008401	0.358226	0.412048	2.799999	30.88009	
2.839999	-0.76135	0.008526	0.363567	0.416761	2.839999	31.33129	
2.879999	-0.77136	0.00865	0.368913	0.421445	2.879999	31.78244	
2.919999	-0.78135	0.008775	0.374263	0.4261	2.919999	32.23354	

2.959999	-0.79131	0.0089	0.379617	0.430726	2.959999	32.68458	
2.999999	-0.80126	0.009024	0.384976	0.435324	2.999999	33.13557	
3.039999	-0.81118	0.009149	0.390339	0.439894	3.039999	33.5865	
3.079999	-0.82108	0.009274	0.395706	0.444436	3.079999	34.03738	
3.119999	-0.83096	0.009399	0.401078	0.448949	3.119999	34.4882	
3.159999	-0.84082	0.009524	0.406454	0.453436	3.159999	34.93896	
3.199999	-0.85065	0.009649	0.411834	0.457894	3.199999	35.38967	
3.239999	-0.86047	0.009774	0.417219	0.462326	3.239999	35.84031	
3.279999	-0.87026	0.009899	0.422607	0.46673	3.279999	36.29091	
3.319999	-0.88003	0.010024	0.428	0.471107	3.319999	36.74144	
3.359999	-0.88978	0.010149	0.433397	0.475457	3.359999	37.19191	
3.399999	-0.89951	0.010274	0.438798	0.47978	3.399999	37.64233	
3.439999	-0.90921	0.010399	0.444203	0.484077	3.439999	38.09269	
3.479999	-0.9189	0.010524	0.449612	0.488348	3.479999	38.54299	
3.519999	-0.92856	0.010649	0.455025	0.492592	3.519999	38.99323	
3.559999	-0.9382	0.010775	0.460441	0.496811	3.559999	39.44342	
3.599999	-0.94783	0.0109	0.465862	0.501003	3.599999	39.89354	
3.639999	-0.95743	0.011025	0.471287	0.50517	3.639999	40.34361	
3.679999	-0.96701	0.011151	0.476715	0.509311	3.679999	40.79362	
3.719999	-0.97657	0.011276	0.482148	0.513427	3.719999	41.24357	
3.759998	-0.98611	0.011401	0.487584	0.517517	3.759998	41.69347	
3.799998	-0.99563	0.011527	0.493024	0.521582	3.799998	42.1433	
3.839998	-1.00512	0.011652	0.498467	0.525623	3.839998	42.59308	
3.879998	-1.0146	0.011778	0.503914	0.529638	3.879998	43.0428	
3.919998	-1.02406	0.011903	0.509365	0.533629	3.919998	43.49247	
3.959998	-1.0335	0.012029	0.51482	0.537596	3.959998	43.94207	
3.999998	-1.04291	0.012154	0.520278	0.541538	3.999998	44.39162	